

Deterministic processes have limited impacts on foliar fungal endophyte communities along a savanna-forest successional gradient

Mathew Harris¹, Martin Kemler², Bernard Slippers¹, Samantha-Leigh Jamison-Daniels¹, Frederick Witfeld², Monique Botha¹, Dominik Begerow², Andreas Brachmann³, and Michelle Greve¹

¹University of Pretoria

²Ruhr-Universität Bochum

³Ludwig-Maximilians-Universität München

July 6, 2022

Abstract

Patterns and drivers of succession provide insight into the mechanisms that govern community assembly and are indicators of community resilience and stability but are still poorly understood in microbial communities. We assessed whether the successional trends of woody vegetation are mirrored by foliar fungal endophyte communities of three tree species that are abundant across the woody successional gradient using a total amplicon sequencing approach. Additionally, we test the relative contribution of host identity, abiotic predictors, biotic factors, and spatial distance between sites in predicting community composition and species richness of endophyte communities. Unlike the woody community, endophyte communities showed no consistent evidence of deterministic successional trends. Host identity was the most important factor structuring fungal endophyte community composition. Spatial distance played some role in explaining differences in community composition, but the effects of this and other environmental variables were small and not consistent between different host species. Much of the variation in endophyte composition remained unexplained. Host identity was most important in predicting endophyte richness. Although endophyte communities showed no deterministic succession, community assembly was most strongly influenced by host identity and spatial distance.

Title: Deterministic processes have limited impacts on foliar fungal endophyte communities along a savanna-forest successional gradient

Mathew Andrew Harris^{1,2}; Martin Kemler³; Bernard Slippers^{2,4}; Samantha-Leigh Jamison-Daniels¹; Frederick Witfeld³; Monique Botha¹; Dominik Begerow³; Andreas Brachmann⁵ & Michelle Greve^{1,2}

1 – Department of Plant and Soil Science, University of Pretoria, Private Bag X20, Pretoria, 0002, South Africa

2 – Forestry and Agricultural Biotechnology Institute (FABI), University of Pretoria, Private Bag X20, Pretoria, 0002, South Africa

3 – Evolution of Plants and Fungi, Ruhr-Universität Bochum, Universitätsstr. 150, 44801 Bochum, Germany

4 – Department of Biochemistry, Genetics and Microbiology, University of Pretoria, Private Bag X20, Pretoria, 0002, South Africa

5 – Genetics, Ludwig-Maximilians-Universität München, Großhaderner Str. 2-4 82152 Planegg-Martinsried München, Germany

Corresponding author’s contact details

e-mail: mathew.harris@fabu.up.ac.za

Address: Department of Plant and Soil Sciences, University of Pretoria, Private Bag X20, Pretoria, 0002, South Africa

phone: +27 12 420 2487

Running title: Decoupled plant-endophyte successional trends.

Significance statement (which places this work in the top 10%): This is the first test of whether endophyte communities follow the successional trends displayed by a woody community undergoing deterministic succession.

Summary

Patterns and drivers of succession provide insight into the mechanisms that govern community assembly and are indicators of community resilience and stability, but are still poorly understood in microbial communities. We assessed whether the successional trends of woody vegetation are mirrored by foliar fungal endophyte communities of three tree species that are abundant across the woody successional gradient using a total amplicon sequencing approach. Additionally, we test the relative contribution of host identity, abiotic predictors, biotic factors, and spatial distance between sites in predicting community composition and species richness of endophyte communities. Unlike the woody community, endophyte communities showed no consistent evidence of deterministic successional trends. Host identity was the most important factors structuring fungal endophyte community composition. Spatial distance played some role in explaining differences in community composition, but effects of this and other environmental variables were small and not consistent between different host species. Much of the variation in endophyte composition remained unexplained. Host identity was most important in predicting endophyte richness. Although endophyte communities showed no deterministic succession, community assembly was most strongly influenced by host identity and spatial distance.

Key words

Community assembly, Ecology, Fungi, Microbiome, Succession, Symbiosis, Trees

Introduction

Understanding the relative contribution of deterministic and stochastic processes during community assembly represents a major unresolved issue in microbial ecology (Dini-Andreote *et al.*, 2015; Antwis *et al.*, 2017; Zhou and Ning, 2017; Tripathi *et al.*, 2018). Unravelling the mechanisms that drive shifts in community composition is crucial for understanding the functions and ultimately the ecosystem processes that microbial communities are able to deliver (Salles *et al.*, 2009; Crowther *et al.*, 2014; Laforest-Lapointe *et al.*, 2017). Succession, the study of how biological communities reorganise through time after disturbances (Johnson, 1979; Chang and Turner, 2019), provides valuable insight into community assembly (Chang and HilleRisLambers, 2016). Understanding succession in microbial communities can, therefore, help to disentangle the relative importance of deterministic and stochastic processes on community assembly (Tripathi *et al.*, 2018).

The composition of communities during succession can be driven by deterministic mechanisms (environmental conditions and biotic filters) and stochastic processes (dispersal, drift, speciation and priority effects) (Hubbell, 2001; Vellend, 2010; Nemergut *et al.*, 2013; Chang and HilleRisLambers, 2016). For microbial communities it has been hypothesised that during the early stages of succession, communities are largely governed by stochastic events, and only as succession continues does the relative strength of deterministic processes begin to cause directional changes in community composition (Dini-Andreote *et al.*, 2015,

2016). However, if stochastic processes such as dispersal are not limiting, they may overpower deterministic processes and the system remains ecologically neutral (Hubbell, 2001).

Studies on fungal succession have mostly focused on belowground soil and root-associated communities (Blaalid *et al.*, 2012; Brown and Jumpponen, 2014; Davey *et al.*, 2015; Dini-Andreote *et al.*, 2016; Dong *et al.*, 2016; Turner *et al.*, 2019). These studies have repeatedly shown that fungal community assembly during later successional stages is deterministic (Brown and Jumpponen, 2014; Dini-Andreote *et al.*, 2016; Gao *et al.*, 2019; Turner *et al.*, 2019). Both abiotic (e.g. carbon, nitrogen, pH and phosphorus) and biotic (e.g. plant species richness, composition, root exudates and microbial competition) factors are drivers of belowground microbial community composition patterns (Dini-Andreote *et al.*, 2016; Dong *et al.*, 2016; Turner *et al.*, 2019). While some evidence exists for both linked (Davey *et al.*, 2015) and decoupled (Turner *et al.*, 2019) trends of succession between plant communities and belowground fungal communities, generalisable patterns across different fungal guilds are still unclear.

Disentangling the relative influence that host-associated and environmental factors play in determining host-endophyte composition represents a major theme in endophyte ecology (Antwis *et al.*, 2017; Harrison and Griffin, 2020), as these fungi are an important aspect in plant performance and health (Compant *et al.*, 2019). Host identity appears to be one of the most important drivers of foliar fungal endophyte community composition (Terhonen *et al.*, 2019). Host-specific defence mechanisms, the host-specific production of various enzymes, secondary metabolites and concentration differences of nutrients and molecules within their leaves directly influence the composition of foliar fungal endophyte communities (Kembel and Mueller, 2014; Cordovez *et al.*, 2019; Darlison *et al.*, 2019; Tellez *et al.*, 2020). Geographic distance is another important factor influencing community composition, with community dissimilarity generally increasing with geographic distance due to dispersal limitation (Soininen *et al.*, 2007). Evidence for distance decay in foliar fungal endophytes is mixed. Some evidence points to the absence of such a relationship at both small (Cordier *et al.*, 2012; Oono *et al.*, 2017) and large scales (Vincent *et al.*, 2016; Barge *et al.*, 2019; U'Ren *et al.*, 2019), while other studies show evidence of distance decay, especially for rare foliar fungal endophyte taxa (Vaz *et al.*, 2014; David *et al.*, 2016; Koide *et al.*, 2017; Oono *et al.*, 2017). Climate, a major factor driving the community composition of plants, seems to have less influence on determining the composition of fungal endophytes, particularly at fine scales (Compant *et al.*, 2010; Santoyo *et al.*, 2017). Water availability in particular affects some fungal endophytes' ability to germinate and persist (Arnold, 2007; Peay *et al.*, 2016). The effect of climate on endophyte composition may also be indirect: through its effect on host composition and physiology fungal endophyte community composition may be affected (Compant *et al.*, 2010; Terhonen *et al.*, 2019).

Plant community composition structures soil and root-associated fungal communities due to the strong biotic filter imposed by plant hosts (Carney and Matson, 2006; Hausmann and Hawkes, 2009; Hoch *et al.*, 2019; Hu *et al.*, 2019). However, how plant community composition structures fungal endophyte composition is yet to be directly assessed (Griffin and Carson, 2018; Griffin *et al.*, 2019). Therefore, directly linking plant community composition to foliar fungal endophyte composition may help to disentangle the factors responsible for structuring these fungal communities and ultimately help to understand how this scales-up to mediate plant microbial ecosystem functioning relationships (Laforest-Lapointe *et al.*, 2017; Griffin *et al.*, 2019; Harrison and Griffin, 2020).

In contrast to community composition, species richness considers only the number of species. A number of factors affect foliar fungal endophyte species richness (Arnold and Lutzoni, 2007; Unterseher *et al.*, 2007; Lau *et al.*, 2013; Griffin *et al.*, 2019; Harrison and Griffin, 2020). Host identity appears to be one of the most important drivers (Lau *et al.*, 2013; Peay *et al.*, 2016; U'Ren *et al.*, 2019; Yao *et al.*, 2019). Other factors, including host age or height, microhabitat (e.g., moisture, light intensity and temperature) and plant richness, can also shape foliar fungal endophyte richness (Bernstein and Carroll, 1977; Unterseher *et al.*, 2007; Zimmerman and Vitousek, 2012; Scholtysik *et al.*, 2013; Oono *et al.*, 2015; Griffin *et al.*, 2019). Yet, few studies have simultaneously assessed how multiple factors drive patterns of foliar fungal endophyte richness.

The aim of this study is to assess patterns and drivers of foliar fungal endophyte community assembly in a system undergoing deterministic vegetation succession. Our objectives were to a) test whether foliar fungal endophyte communities follow a deterministic successional trend as displayed by the woody tree communities in the same study system; and b) determine the relative contribution of host species identity, surrounding tree composition, geographic distance and abiotic variables to endophyte composition and richness.

Experimental procedures

Study area

The study was conducted at Buffelskloof Private Nature Reserve (250 19' 22.21" S, 300 29' 15.41" E), Mpumalanga, South Africa. Buffelskloof supports vegetation types from three biomes: Afromontane forest, mid-altitude savanna and montane grassland (Mucina and Rutherford, 2011). The reserve is approximately 1500 ha in size with an altitudinal variation of 1000 to 1800 m (Buffelskloof Nature Reserve, 2019). Samples were collected between 1314 and 1477 m elevation.

Within the grassland and savanna vegetation on Buffelskloof, bush clumps (BCs) have been establishing (Jamison-Daniels *et al.*, 2021). A BC is defined as an association of one or more trees (> 1.2 m) with tree saplings (< 1.2 m) growing beneath the canopy, that are separated from each other by grassy vegetation (Jamison-Daniels *et al.*, 2021). A space-for-time substitution study, in which the woody vegetation of 40 BCs ranging in size from 10.05 m² to 1342.99 m² (mean = 266.41 m², SE = 347.75 m²) was characterised, where it was found that the formation of these BCs follows a deterministic trend of succession in woody species, with a turnover from open-habitat trees to shade-tolerant trees as the BCs increase in size (Jamison-Daniels *et al.*, 2021). These directional changes in tree community composition are driven by directional changes in microclimatic conditions (temperature, soil moisture and humidity) and light availability as BCs increase in size (Jamison-Daniels *et al.*, 2021).

Sampling

Endophyte communities were sampled from the same BCs that were surveyed by Jamison-Daniels *et al.*, (2021). Three tree species that are widely distributed and common across BCs of different sizes (and thus different successional stages) (Jamison-Daniels *et al.*, 2021) were selected: *Euclea crispa* subsp. *ovata* (Burch.) F.White, *Searsia chirindensis* (Baker f.) Moffett and *Canthium inerme* Kuntze. Trees of these species were present in 38 of the 40 BCs.

Field sampling took place between 20 and 23 November 2018, to characterise the summer leaf endophyte communities. In every BC containing the host species, up to four trees (> 1.2 m) per host species were selected. The coordinates of each tree that was sampled were recorded using a Garmin Etrex GPS (Garmin Ltd., Olathe, KS, USA) and the height of each tree was estimated using a 1.2m dowel stick. All hosts did not occur in all BCs, and BCs did not always contain four individuals from each host. For BCs that had more than four individuals of a host, the first four individuals that were encountered were sampled. *E. crispa*, *C. inerme* and *S. chirindensis* trees were present in 27, 21 and 28 of the BCs, and 84, 50 and 55 individual trees were sampled from each host species, respectively. From each tree, five leaves from each of the four cardinal directions were collected half-way between the highest and lowest leaves. Only fully unrolled leaves that had reached maturity and had no visible signs of infection or insect damage were selected. Leaves were stored in envelopes within a cooler box, and processed within 8 hours of collection.

Microclimatic conditions within each BC were characterised to determine whether differences in microclimate influenced endophyte community composition and richness (see Jamison-Daniels *et al.*, 2021 for full details). Light intensity, temperature (average, minimum, maximum and standard deviation measured over 130 days), relative humidity and soil moisture were all previously measured and subsequently extracted (from Jamison-Daniels *et al.*, 2021).

The species richness (count of tree species per BC), Shannon-Wiener diversity, species composition and the tree basal area of all trees >1.2 meters per BC were obtained from Jamison-Daniels *et al.*, (2021). Only trees >1.2 meters were considered, as these represent established vegetation within the BCs. Additionally,

proximity of each sampled tree to the large indigenous forest which acted as a source population for most BC trees (Jamison-Daniels *et al.*, 2021) was calculated to represent a proxy for potential endophyte inoculum pressure. Within Google Earth Pro (v7.3.2.5776) a polygon was drawn around the indigenous forest at Buffelskloof Nature Reserve and subsequently extracted as a shape file. The distance from each tree sampled to the edge of the indigenous forest within the reserve was obtained using the points to polygon function in Esri, ArcMap (Esri, Redlands, CA, USA).

Leaves underwent surface washing (Arnold and Herre, 2003) to reduce and possibly eliminate the epiphytic burden from each sample. All leaves sampled from one tree were washed together successively in 70% EtOH (30 seconds), 2% NaOCl (60 seconds), 70% EtOH (60 seconds) and autoclaved dH₂O (60 seconds). The leaves were then placed on paper towel and left to dry. Leaf disks from the dried leaves were cut using a 6 mm cork-borer (sterilised between each sample) and subsequently stored on silica gel in falcon tubes until DNA extraction. For *E. crista* and *C. inermis* 12-18 leaf disks were cut per leaf, while 22-25 disks were cut from larger *S. chirindensis* leaves.

Homogenisation, DNA extraction and sequencing

Prior to homogenisation of each sample, a metal cylinder and ball bearing were sterilised in a 6% sodium hypochlorite (NaOCl) solution for three minutes, followed by immersion in 70% ethanol (EtOH) for one minute. All leaf disks from one sample were homogenised together in the metal cylinder fastened to a Retsch® MM2000 laboratory mixer mill (Retsch® GmbH, Haan, Germany) for one minute at 70% of the maximum oscillation frequency and subsequently stored at -200C until DNA extraction. DNA of 60 mg dried homogenised leaf material per sample was extracted using the my-Budget plant DNA extraction kit (Bio-Budget Technologies GmbH, Krefeld, Germany) following manufacturer's instructions. The ITS region of all extracted samples was amplified using the ITS1-F (Gardes and Bruns, 1993) and ITS4 (White *et al.*, 1990) primer combination and was subsequently visualised on a 0.8% agarose gel with a 100bp ladder, together with positive and negative controls to ensure that we had indeed extracted fungal DNA without contamination.

Illumina amplicon library preparation utilised a nested PCR approach (Unterseher *et al.*, 2016) (Supporting Information Figure 1, Methods S1). The ITS region was amplified using the ITS1-F (Gardes and Bruns, 1993) and ITS4 (White *et al.*, 1990) primer combination. The final amplicons were subsequently visualised on a 0.8% agarose gel with added ethidium bromide (Supporting Information Methods S1). Concentration of DNA for pooling was quantified from the band intensity of the gel images using ImageJ (Schneider *et al.*, 2012) (Supporting Information Methods S1). The amplicon pools were cleaned using the CleanPCR magnetic bead kit (CleanNA, Waddinxveen, Netherlands) (Supporting Information Methods S2). In total three samples were lost during homogenisation or due to low DNA concentration resulting in 186 samples that were sent for sequencing.

The final pooled amplicon library was sequenced at the Genomics Service Unit of the Ludwig Maximilians University (LMU) Biocenter on an Illumina MiSeq® sequencer (Illumina Inc., San Diego, CA, USA) using the MiSeq® Reagent Kit v3 Chemistry, for 2 × 250 bp paired-end sequencing. All raw sequences were deposited on the NCBI portal under the following accession codes: BioProject – PRJNA674320; BioSample (SAMN16634781 – SAMN16634966).

Bioinformatics

Bioinformatic processing was performed in QIIME1 (Caporaso *et al.*, 2010) and QIIME 2 (Bolyen *et al.*, 2019). Samples were demultiplexed according to their unique tag/index combinations (Supporting Information Table S1), which were removed during the process, together with the primer sequences (Supporting Information Table S2). For subsequent analyses only forward reads were used, as reverse reads often suffer from a lower PHRED quality and due to length differences within the ITS gene region, which often prevents merging of both reads. Three *S. chirindensis* samples which contained less than 10 000 sequences were removed prior to downstream bioinformatic analyses. The raw sequences from the remaining 183 samples were subsequently passed through deblur (Amir *et al.*, 2017) implemented in the QIIME 2 pipeline, which assigns

raw sequence reads to Amplicon Sequence Variants (ASVs). Reads were trimmed at 180 bp. The UNITE database was used as reference sequences (version 8; 020219) (<https://unite.ut.ee/>). Only ASVs which were classified as belonging to the Kingdom Fungi were retained. Fungal ASVs were written into a feature table, which was used for subsequent downstream analyses. The full ASV feature table and all metadata relating to the manuscript can be obtained from figshare: 10.6084/m9.figshare.14518200 – ASV feature table and 10.6084/m9.figshare.14518218 – metadata.

Analyses

All analyses were performed on the full, unrarified, ASV table. This was done for two reasons. First, rarefying the ASV table to the smallest sample size to account for differences in library sizes between samples made no difference to the interpretation of the results (results not shown). Second, from a statistical point of view, rarefaction is inept for the comparison of relative abundances (McMurdie and Holmes, 2014; Willis, 2019).

As predictor variables that are highly correlated can lead to spurious effects on analyses, all continuous predictor variables were tested for multi-collinearity prior to analyses (Supporting Information Table S3). When two variables were highly correlated, i.e. $r > 0.75$, one of these variables was removed (Supporting Information Table S3). Bush clump area, bush clump tree basal area and bush clump tree species richness were highly correlated. Therefore, only bush clump tree basal area was retained for analyses on endophyte composition and richness, as bush clump tree basal area gives a good representation of available woody tree host density within individual BCs. Bush clump area was only retained in the analyses on successional trends, as BC area was a good proxy for BC maturity and woody vegetation successional stage (Jamison-Daniels *et al.*, 2021).

Successional trend

A two-step approach was used to determine the processes of endophyte community assembly and to evaluate whether these fungal communities followed the deterministic successional trend observed for their woody tree hosts (Jamison-Daniels *et al.*, 2021).

First, we assessed whether the amount of variation in community composition of our actual endophyte assemblages, explained by a number of predictor variables, was greater than (implying deterministic processes), less than (implying deterministic processes) or not different (implying stochastic processes) to what could be expected in randomly generated communities (Dini-Andreote *et al.*, 2015). Initially, a canonical correspondence analysis (CCA) was conducted to test which factors, namely host identity, spatial distance between sampled trees, abiotic conditions (maximum and minimum temperature, temperature standard deviation, light intensity, soil moisture measured per BC) and biotic factors (tree basal area per BC, tree height and distance to forest edge), best explained fungal community composition. Weighted linear regression was performed on the constraining variables (Ter Braak, 1986). To incorporate the effect of spatial distance on community composition, a weighted principal coordinates of neighbourhood matrix (PCNM) analysis (created using the geographic coordinates of every sampled tree) was performed. This analysis transforms coordinates to a rectangular distance matrix that is acceptable for constrained ordination techniques like CCA (Borcard and Legendre, 2002; Legendre and Borcard, 2008). Backward and forward stepwise permutation tests, for 1000 permutations, were used to determine the best fitting model, based on the model with the lowest AIC score (Venables and Ripley, 2002). Rare species contribute heavily to the chi-squared distance used to plot site and species scores in CCA analysis (Legendre and Legendre, 2012). Therefore, to reduce the spurious effects of rare species within the CCA, an eigenvalue decomposition approach was used to determine in how many samples a particular ASV must have occurred for it to be retained when performing the final ordination (Legendre and Legendre, 2012). When a considerable drop in inertia for one of the first five eigenvalues is detected, it indicates in how many samples an ASV must have occurred to be retained for the final CCA analysis (Supporting Information Figure S2) (Legendre and Legendre, 2012). The eigenvalue decomposition detected a drop in the inertia for the third eigenvalue after dropping ASVs that occurred in less than 20 samples (Supporting Information Figure S2). Therefore, all ASVs ($n= 4905$) that occurred in 20 samples or less were removed before the final CCA analysis ($N= 411$) and the randomisations.

The amount of variation explained by the best model with the retained predictor variables (host identity, light intensity and four spatial eigenvectors (spatial eigenvector 1, spatial eigenvector 2, spatial eigenvector 13 and spatial eigenvector 17)) was calculated. Then, 10 000 randomly assembled abundance-weighted community data matrices were constructed based on the true community identity, richness and abundance (Greve *et al.* , 2008). The randomly assembled matrices conserved species richness per sample as observed in the true community, and set the probability of species being selected proportional to ASV read abundance (Gotelli and Graves, 1996; Gotelli, 2000). For each of the 10 000 randomly assembled communities, a new CCA was performed with the predictors from the best CCA model (see above). The 2.5% and 97.5% quantiles of the percentage variation explained by the CCAs were calculated for the 10 000 random communities, and it was assessed whether the percentage variation explained by the CCA of the true community was larger than ($> 97.5\%$ quantile), smaller than ($< 2.5\%$ quantile) or not significantly different (between 2.5% and 97.5% quantiles) to the percentage variation explained by the randomly generated communities using a z-test (Greve *et al.* , 2008).

Because CCA randomisation analyses provided evidence for deterministic community assembly (see Results), a second analysis was conducted to test whether changes in community composition could be explained by BC size, as BC size increases with tree succession (Jamison-Daniels *et al.* , 2021). A significant directional change in endophyte species composition with BC size would indicate deterministic succession, while no predictable change in composition with BC size would suggest stochastic succession (Dini-Andreote *et al.* , 2015). We tested this for the fungal communities extracted from each of the three host species using two different pair-wise similarity indices (Morisita and Raup-Crick indices). The Morisita index is weighted towards assessing similarity in common taxa, while the Raup-Crick index gives more weight to co-occurring rare taxa (Morisita, 1962; Raup and Crick, 1979). Pair-wise similarity in fungal community composition was calculated between all possible combinations of each of the smallest $\frac{1}{4}$ of the BCs and each of the largest $\frac{3}{4}$ of the BCs for each host species (following Jamison-Daniels *et al.* , 2021), to establish if there was directional change in community composition as BCs increase in size. For each of the largest $\frac{3}{4}$ BCs, similarity values with the smallest $\frac{1}{4}$ of the BCs were averaged. Generalised linear mixed effects models (GLMMs) were used to model the effect of BC area (of the largest $\frac{3}{4}$ BCs) on foliar fungal endophyte community similarity for both Raup-Crick and Morisita similarity indices. BC identity of the largest $\frac{3}{4}$ BCs was included as a random effect in the model (McCulloch, 1997). Since the similarity values for both the indices are scaled between 0-1, models were fitted using a binomial distribution and logit link function (Zuur *et al.* , 2009). If the foliar fungal community similarity between the largest and the smallest BCs decreased or increased significantly with BC area of the largest BCs, it was interpreted as an indication of deterministic succession, while no relationship was taken as an indication of communities being ecologically neutral and primarily governed by stochastic processes that structure community composition (Hubbell, 2001; Dini-Andreote *et al.* , 2015; Jamison-Daniels *et al.* , 2021).

Drivers of assemblage composition

We tested which factors, namely host identity, spatial distance between sampled trees, abiotic conditions (maximum and minimum temperature, temperature standard deviation, light intensity and soil moisture measured per BC) and biotic factors (tree basal area per BC, tree height, distance to forest edge and the tree composition per BC), affected fungal endophyte community composition.

To do this we used a generalized dissimilarity modelling (GDM) approach. This flexible approach enabled us to simultaneously incorporate categorical (i.e. host), linear (e.g. climate), compositional (i.e. tree compositional dissimilarity), and spatial data (i.e. geographic distance) predictors of endophyte composition into one analysis (Ferrier *et al.* , 2007). GDM is a nonlinear extension of matrix regression, which has specifically been designed to deal with two types of nonlinearity commonly encountered in biological data: 1) the curvilinear relationship between ecological or spatial separation and the observed compositional dissimilarity, and 2) non-stationarity, i.e. differences in the rate of compositional turnover along environmental or spatial gradients (Ferrier *et al.* , 2007; Fitzpatrick *et al.* , 2013).

The default of three I-spline basis functions (knots) per predictor variable was used in all GDM analyses

(Ferrier *et al.* , 2007); and backwards selection was used to determine how many variables to retain in each of the final models (Williams *et al.* , 2012). The sum of the coefficients per I-spline represents the maximum amount of variation explained by a particular variable, and can be used to determine variable importance (Ferrier *et al.* , 2007). Since host identity explained the majority of the endophyte compositional dissimilarity for the full dataset (see Results), we repeated the GDM analyses per individual host species using the same approach as above.

Drivers of endophyte richness

To test the effects of host identity (i.e. host species), abiotic variables (maximum and minimum temperatures, temperature standard deviation, light intensity, relative humidity and soil moisture) and biotic variables (tree basal area per BC, tree height and distance to the forest edge) on ASV richness, GLMMs, with a Poisson distribution and a log-link function were used (Zuur *et al.* , 2009). Interaction terms between host identity and all abiotic and biotic variables were fit within the model and BC identity was included in the model as a random variable (McCulloch, 1997). Best subset modelling, based on the lowest AIC-value, was used to assess which predictor variables from the global model should be retained (Burnham and Anderson, 2002). The model was overdispersed, therefore overdispersion was corrected by employing the observation-level random effects approach (Lawson *et al.* , 1999; Elston *et al.* , 2001). Marginal and conditional R^2 -values were calculated (Nakagawa *et al.* , 2017).

Statistical software

All analyses were conducted in R, v3.6.0 (R Core Team, 2019) using the packages *vegan* v2.5-6 (Oksanen *et al.* , 2019), *lme4* v1.1-23 (Bates *et al.* , 2015), *MuMIn* v1.43.17 (Barton, 2019), *car* v3.0-9, *gdm* v1.4.2.2. (Fitzpatrick *et al.* , 2021) , and *effects* v4.2-0 (Fox, 2003; Fox and Weisberg, 2019).

Results

The 183 leaf samples from three native host species yielded 7 355 098 demultiplexed sequences. On average there were 40 192 sequences per sample, the highest number of sequences from one sample was 103 727 and the lowest 12 533 (Supporting Information Table S5). In total 5 326 unique ASVs were recovered with an average of 193.80 ± 101.82 ASVs per sample. The number of ASVs recovered from *C. inermis* , *E. crispa* and *S. chirindensis* was 1 473, 2 221 and 1 847, respectively.

Successional trend

Variables retained in the CCA (host identity, spatial distance between samples and light intensity) explained 11.76% of the variation of the actual endophyte community. In contrast, only between 4.27% - 4.75% of the variation in ASV composition of randomly generated communities could be explained by these same predictor variables. Therefore, the predictor variables predicted significantly (2.5 times) more variation in community composition of the true than the randomly generated communities (z -value = 56.812, p -value = $<16 \times 10^{-16}$). This indicates that communities are non-randomly assembled and that deterministic selective forces structure fungal endophyte communities.

The similarity of foliar fungal communities showed no consistent trends across the gradient of BC size (Table 1). The average Raup-Crick similarity decreased significantly with increasing BC area for endophyte communities of *E. crispa* (Figure 1 & Table 1), as expected under deterministic succession. However, no significant trends in community similarity with BC area could be observed for rare communities of the other two host species (Table 1). Similarly, the Morisita similarity index showed no host-specific successional trends across BC area (Table 1). Therefore, five of the six analyses showed no evidence of deterministic succession.

Assemblage composition

The final GDM for the full dataset retained two host species (*E. crispa* and *S. chirindensis*), tree basal area, distance to the indigenous forest edge, geographic distance between samples, light intensity and tree compositional dissimilarity as predictor variables (Figure 2). The final model for the full dataset was significant (Null Deviance = 691.996, GDM Deviance = 491.635, p -value = 0.000001), as were all the predictor

variables, except light intensity (Figure 2; Table 2). The final model explained 28.95% of the deviance in the turnover of foliar fungal endophyte community composition. With host identity alone explaining ~23% of this variance. Of the remaining predictors distance to the forest and the difference in tree basal area were the most important, both having variable rates of turnover along the gradients and displaying the quickest turnover at short distances from the forest and small differences in tree basal area (Figure 2c & d). The rate of turnover in fungal endophyte composition along the tree compositional dissimilarity gradient was the steepest of any of the gradients; with the maximum magnitude of turnover reached when the difference in surrounding tree composition was only slightly dissimilar, i.e. samples not from the same BC (Figure 2f).

When GDMs were run per host species, different predictors were retained as important in explaining the composition of endophytes within different hosts. Only geographic distance between samples was retained and significant in all models (Figure 2 & Figure 3). The final GDM model for *E. crispa* was significant (Null Deviance = 147.69, GDM Deviance = 122.65, p -value = 0.000001) and explained 16.96% of the deviance in turnover of foliar fungal endophyte community composition. It retained five variables (Supporting Information Table S5), of which distance to the forest edge, geographic distance between samples and the difference in tree basal area were significant (Figure 3a-c). Distance to the forest was the most important predictor, with the rate of turnover being approximately linear (Figure 3a). The rate of turnover in endophyte composition along the geographic distance and tree basal area gradients were non-linear with the highest rates of turnover occurring at larger geographic separation between samples and small differences in tree basal area (Figure 3b & 3c).

For *C. inerme*, the final GDM model retained five predictor variables (Supporting Information Table S6); however, only minimum temperature and geographic distance between samples were significant (Figure 3d & 3e). Minimum temperature was the most important predictor with both larger differences in minimum temperature and geographic distance between samples resulting in higher rates of turnover. The magnitude and rates of turnover in endophyte composition was almost non-existent at low differences in minimum temperatures and short geographic distances between samples but increased dramatically at the higher end of these gradients (Figure 3d & 3e). The final model was significant (Null Deviance = 52.85, GDM Deviance = 45.92, p -value = 0.00503), and explained 13.12% of the deviance in turnover of foliar fungal endophyte community composition.

The final GDM model for *S. chirindensis* fungal communities retained five predictor variables (Supporting Information Table S7); but the only retained variable which was significant was geographic distance between samples (Figure 3e). The rate of turnover in endophyte composition increased with distance between samples; this increase was most rapid at short geographic distances (Figure 3e). The final *S. chirindensis* GDM model was significant (Null Deviance = 42.78, GDM Deviance = 38.622, p -value = 0.000001), and explained 9.52% of the deviance in turnover of foliar fungal endophyte community composition.

ASV richness

Host was the most important predictor of ASV fungal richness (Table 3), with one host – *S. chirindensis* – consistently supporting fewer ASVs (mean = 83.94 ± 52.79) than *C. inerme* (mean = 218.5 ± 95.8) and *E. crispa* (mean = 245.85 ± 70.82) (Figure 4). The best subset GLMM model testing the effect of abiotic, biotic and spatial factors on ASV fungal endophyte richness retained light intensity; the interaction between host species and tree height and the interaction between host species and maximum temperature. Both the interactions between host and maximum temperature, and between host and tree height were significant (Figure 4, Table 3). Fungal ASV richness increased as light intensity decreased (Figure 4a). ASV endophyte richness decreased in *S. chirindensis* individuals that experienced higher maximum temperatures (Figure 4b) but was unaffected by temperature differences experienced by individuals of *C. inerme* and *E. crispa* (Figure 4b). Additionally, ASV endophyte richness decreased with tree height for *S. chirindensis* but did not change with tree height in *C. inerme* and *E. crispa* (Figure 4c). Fixed effects explained 63% of variation in ASV richness (Table 3).

Discussion

Foliar fungal endophyte communities did not display consistent patterns of deterministic succession within individual host species, despite the surrounding plant community undergoing deterministic successional changes. Instead, host species identity was the most important factor shaping community composition and richness patterns for endophytes. The strong effect imposed by the host potentially limited the ability to observe consistent deterministic succession patterns across the successional gradient, i.e. BC size and its associated changes in abiotic conditions.

Successional patterns

Across the common and rare endophyte assemblages of three host tree species, evidence for deterministic succession was only found for the rare endophyte communities of one of the three woody host species (*E. crisper*) and in none of the three assemblages of common endophyte species. Therefore, we find no consistent evidence of deterministic succession of endophyte communities within individual hosts across a woody-plant successional gradient. Under a scenario of deterministic succession, similar directional shifts in fungal endophyte community composition would have been observed since abiotic environmental conditions changed fairly consistently with BC size (Jamison-Daniels *et al.*, 2021). The lack of consistent directional changes in endophyte composition with BC size, but significant effects of the host identity on species composition, suggest that the effect of the plant host on endophyte composition trumps the impact that the consistent changes of environmental variables has on fungal endophyte communities. A large percentage of variation in the community assembly models remain unexplained, suggesting that unmeasured or random processes may contribute to the assembly of endophyte communities (Hubbell, 2001).

The lack of strong evidence for deterministic succession within endophyte communities does not preclude succession of endophyte communities of all trees per bush clump. Our study specifically assessed changes in endophyte composition within single host species across a tree successional gradient. Because tree composition changed in a directional manner with BC size (Jamison-Daniels *et al.*, 2021), and host identity is an important determinant of endophyte composition and richness (Liu *et al.*, 2019; U'Ren *et al.*, 2019; Harrison and Griffin, 2020) (Figure 2; Figure 4; Table 1), endophyte communities of entire BCs (i.e. across different tree species or within tree genera or families) could show stronger evidence of deterministic changes in endophyte composition. As a certain suite of tree species are associated with a specific successional stage in this system (Jamison-Daniels *et al.*, 2021), their associated endophyte communities may also be associated with the associated tree successional stage. Under this scenario, the succession of endophyte communities would be driven by changes in host species composition with successional stage, and not changes in environmental factors with successional stage.

Drivers of community composition

Although no consistent evidence of deterministic succession in endophyte communities was observed; there was some evidence of deterministic processes shaping the community assembly of fungal endophyte communities. Host identity had the largest impact on endophyte community composition (Figure xx, Table xx). Host identity alone explained ~23% of the total ~29% variation in turnover of endophyte composition explained by the final model. Host identity is known to be a major determinant of the composition of foliar fungal endophyte communities (Christian *et al.*, 2016; David *et al.*, 2016; Vincent *et al.*, 2016; Liu *et al.*, 2019; Yao *et al.*, 2019). The specificity of endophytes to a particular host or group of taxonomically related hosts ranges from extremely narrow to relatively wide (Arnold and Lutzoni, 2007; Pölme *et al.*, 2018). This may explain why, even though host identity was the important factor structuring foliar fungal endophyte communities, the identity of only two host species was retained in the final model, with *C. inermis* consistently being removed as one of the variables which explained the least turnover in endophyte community composition.

Geographic distance between samples was the only significant variable consistently retained in every GDM (i.e. the full dataset and each of the three hosts individually) (Figure 2 & Figure 3)). This indicates that the similarity in composition of foliar fungal endophytes decays with increasing distance, a pattern commonly observed in ecological communities (Nekola and White, 1999; Soininen *et al.*, 2007). Indeed, others have

shown that foliar fungal endophytes are dispersal limited over distances from 10km to >100km (David *et al.* , 2016; Koide *et al.* , 2017; Oono *et al.* , 2017). However, here we show that even over a shorter spatial distance, i.e. <1500m, there is consistent evidence for distance decay driving turnover in endophyte community composition.

The GDMs performed for each host individually, showed that the factors responsible for the turnover in foliar fungal endophyte community composition within a host is variable as is the rate and magnitude of this turnover (Figure 3, Supporting information Table S5-S7). For example, distance to the indigenous forest, the difference in minimum temperature and geographic separation were the most important variables explaining the turnover in foliar fungal endophyte communities for *E. crispa*, *C. inerme* and *S. chirindensis* respectively, holding all other variables constant in their individual models. Others have shown how the distance to a potential inoculum source, in our case a large indigenous forest, can be important for causing compositional differences in fungal communities (Glassman *et al.* , 2017). It has also been shown how abiotic conditions, like temperature differences, can shape endophyte community composition due to differences in the physiological performance of endophyte species in dealing with the prevailing conditions (Arnold and Herre, 2003; Peay *et al.* , 2016; Unterseher *et al.* , 2016). Such findings highlight how both deterministic (i.e. temperature differences) and stochastic (i.e. distance to an inoculum source or geographic separation) forces can be important factors causing turnover in foliar fungal endophyte communities in the same system but for different hosts.

Endophyte richness

Endophyte richness was most strongly influenced by host identity, with endophyte richness of one host (*S. chirindensis*) being consistently lower than that of the other hosts. In contrast, several abiotic and biotic variables showed inconsistent effects on endophyte richness across the three host species. Reduced endophyte richness at higher light intensity agrees with other studies: high levels of UVB negatively impact endophyte persistence, with high UVB increasing leaf desiccation and/or the activation of the plant's defence responses, ultimately leading to lower endophyte richness (Unterseher *et al.* , 2007, 2012).

Other abiotic effects on endophyte richness differed between host species which may shed some light on how such factors drive endophyte richness in these host species. Lower endophyte richness in *S. chirindensis* occurred in warmer BCs that this species occupied. Temperature differences within BCs may drive differences in chemistry and secondary metabolite production produced by the leaves of *S. chirindensis* that experience higher temperatures (Reich *et al.* , 1999; Veteli *et al.* , 2002), potentially leading to lower ASV richness (Arnold and Herre, 2003; Unterseher *et al.* , 2012, 2013). High temperatures can activate the plant defence response, thereby increasing the production of secondary metabolites with potential anti-microbial properties (Unterseher *et al.* , 2016) which decreases the observed endophyte richness by reducing the number of successful endophyte colonisations. Taller trees are usually older, and older hosts have been shown to support decreased endophyte richness compared to their younger conspecifics (Oono *et al.* , 2015). It has been postulated that older trees invest more in defence mechanisms which resist endophyte colonisation; alternatively established groups of endophytes in older trees outcompete newly arriving endophytes, ultimately leading to lower endophyte richness in older trees (Unterseher *et al.* , 2007; Oono *et al.* , 2015).

Conclusion

Although foliar fungal endophyte community assembly within this natural system was structured by deterministic selective forces (particularly host identity), there was no consistent evidence of deterministic successional trends in endophyte communities. A fair amount of variation in endophyte community composition remained unexplained. This could be due to several reasons. Variables which were not measured here, e.g. competitive interactions between endophyte species (Crowther *et al.* , 2014), could be more important for structuring endophyte communities. Alternatively, stochastic processes of community assembly may have been more important than deterministic processes in structuring foliar fungal endophyte composition (Hubbell, 2001). These stochastic processes include drift, dispersal limitation, speciation and extinction events acting together to shape the observed community composition (Vellend, 2010). Additionally, the scale at

which endophyte communities were quantified may not accurately reflect the scale at which processes of endophyte community assembly are operating and thus ultimately obscure our ability to detect strong patterns (Harrison and Griffin, 2020). Thus, continued work to gain a deeper understanding of how fungal endophyte communities organise themselves through space and across time within natural systems will be essential to appreciate the functions and ultimately the ecosystem services these microbes are able to deliver.

Acknowledgements

The authors thank Buffelskloof Private Nature Reserve, more specifically Mr John Burrows, for providing them with access to the research site, and to Mpumalanga Tourism and Parks Agency for permits to collect plant material. A special thanks to Mandy Messal and Laura Milne for their help during the collection and processing of samples collected. The authors would also like to thank the Tree Protection Co-operative Program (TPCP), the centre of excellence in plant health and biotechnology (CPHB) and the South African National Research Foundation (Grant numbers 98889 and 116333) for funding this research.

Author Contributions

MAH, MK, BS & MG contributed to the design of the research, performance of the research, interpretation of the data and writing of the manuscript. MAH, S-L J-D & MB contributed to the data collection and data analyses of the research. MAH, FW, DB & AB contributed to the performance of the laboratory work, sequencing and bioinformatics of the fungal communities.

Data Availability

All raw Illumina sequence data are available on the NCBI portal under the following accession codes: BioProject – PRJNA674320; BioSample (SAMN16634781 – SAMN16634966). The full ASV feature table can be obtained from figshare: 10.6084/m9.figshare.14518200. All metadata relating to the manuscript can be obtained from figshare: 10.6084/m9.figshare.14518218. All R scripts relating to the project can be available upon request to the corresponding author.

Competing Interests

We know of no conflicts of interest that are associated with this potential publication. Additionally, we declare that there has been no significant financial support for conducting this research that could have influenced its overall outcome. As the corresponding author, I confirm that the entire manuscript with all supplementary materials has been read and approved for submission by all authors named above.

Host species	Similarity index	Fixed effects	Estimate	SE	z-value	p-value	Conditional R
<i>Euclea crispa</i>	Raup-Crick	Area	-1.3301	0.4968	-2.677	7.43X10⁻³	0.2185
<i>Euclea crispa</i>	Morisita	Area	-0.16368	0.1528	-1.071	2.84X10 ⁻¹	0.00133
<i>Canthium inerme</i>	Raup-Crick	Area	-0.09949	0.594	-0.167	8.67X10 ⁻¹	0.0613
<i>Canthium inerme</i>	Morisita	Area	-0.6755	0.3599	-1.877	6.05X10 ⁻²	0.0323
<i>Searsia chirindensis</i>	Raup-Crick	Area	0.0064	0.00051	1.248	2.12X10 ⁻¹	0.1082
<i>Searsia chirindensis</i>	Morisita	Area	0.0001	0.00031	0.323	7.47X10 ⁻¹	5.13X10 ⁻⁴

Table 1. Results from the generalised linear mixed effects models assessing the effect of bush clump area on the similarity of fungal communities. Pair-wise similarity values are calculated for all samples taken from each of the three different host species. Raup-Crick index is the probabilistic similarity of co-occurrences between rare and common ASVs, while the Morisita similarity gives weighting to more abundant ASVs.

Table 2. Relative importance of the retained predictor variables for fungal endophyte compositional turnover for the full dataset. Estimates are determine by summing the coefficients of the I-splines from the GDM. Significant *p*- values of retained predictor variables are indicated in bold.

Gradient	estimate	<i>p</i> -value
<i>Searsia chirindensis</i>	0.364518	< 0.0001
<i>Euclea crispa</i>	0.325869	< 0.0001
Distance to forest edge	0.251447	< 0.0001
Δ Tree basal area	0.234689	< 0.0001
Geographic distance	0.174214	< 0.0001
Δ Tree compositional dissimilarity	0.154513	< 0.0001
Inverse light intensity	0.154031	0.075

Fixed Effects	Df	Chi-square	<i>p</i> -value	Marginal R ²
Host	2	190.7435	< 2.2X10⁻¹⁶	0.63055
Maximum temperature	1	1.5634	2.11X10 ⁻¹	
Tree height	1	3.6442	5.63X10 ⁻²	
Tree species richness	1	2.3464	1.26X10 ⁻¹	
Inverse light intensity	1	4.0973	4.30X10⁻²	
Host X Maximum temperature	2	12.6229	1.82X10⁻³	
Host X Tree height	2	6.1574	4.60X10⁻²	
Random Effects	Variance	Standard deviation	AIC	Conditional R ²
(1 Bush clump ID)	0.03451	0.1867	2078.7	0.98935
(1 Sample ID)	0.13616	0.369		

Table 3. Type II Wald chi-square test of the best subset GLMM assessing which factors affect ASV foliar fungal endophyte richness across a tree successional gradient. The retained variables from the best subset model, and their interactions between each other are fixed effects. Bold *p*-values represent significant effects at *p* > 0.05. Bush clump identity and Sample identity were included as random variables.

References

- Amir, A., McDonald, D., Navas-Molina, J.A., Kopylova, E., Morton, J.T., Zech Xu, Z., et al. (2017) Deblur rapidly resolves single-nucleotide community sequence patterns. *mSystems* **2** : e00191-16.
- Antwis, R.E., Griffiths, S.M., Harrison, X.E., Aranega-Bou, P., Arce, A., Bettridge, A.S., et al. (2017) Fifty important research questions in microbial ecology. *FEMS Microbiol Ecol* **93** : fix044.
- Arnold, A.E. (2007) Understanding the diversity of foliar endophytic fungi: progress, challenges, and frontiers. *Fungal Biol Rev* **21** : 51–66.
- Arnold, A.E. and Herre, E.A. (2003) Canopy cover and leaf age affect colonization by tropical fungal endophytes: ecological pattern and process in *Theobroma cacao* (Malvaceae). *Mycologia* **95** : 388–398.
- Arnold, A.E. and Lutzoni, F. (2007) Diversity and host range of foliar fungal endophytes: are tropical leaves biodiversity hotspots? *Ecology* **88** : 541–549.
- Barge, E.G., Leopold, D.R., Peay, K.G., Newcombe, G., and Busby, P.E. (2019) Differentiating spatial from environmental effects on foliar fungal communities of *Populus trichocarpa*. *J Biogeogr* **46** : 2001–2011.
- Barton, K. (2019) MuMIn: Multi-Model Inference.
- Bates, D., Maechler, M., Bolker, B., and Walker, S. (2015) Fitting linear mixed-effects models using lme4. *J Stat Softw* **67** : 1–48.
- Bernstein, M.E. and Carroll, G.C. (1977) Internal fungi in old-growth Douglas fir foliage. *Can J Bot* **55** : 644–653.

- Blaalid, R., Carlsen, T., Kumar, S., Halvorsen, R., Ugland, K.I., Fontana, G., and Kauserud, H. (2012) Changes in the root-associated fungal communities along a primary succession gradient analysed by 454 pyrosequencing. *Mol Ecol* **21** : 1897–1908.
- Bolyen, E., Rideout, J.R., Dillon, M.R., Bokulich, N.A., Abnet, C.C., Al-Ghalith, G.A., et al. (2019) Reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. *Nat Biotechnol* **37** : 852–857.
- Borcard, D. and Legendre, P. (2002) All-scale spatial analysis of ecological data by means of principal coordinates of neighbour matrices. *Ecol Modell* **153** : 51–68.
- Ter Braak, C.J.F. (1986) Canonical correspondence analysis: a new eigenvector technique for multivariate direct gradient analysis. *Ecology* **67** : 1167–1179.
- Brown, S.P. and Jumpponen, A. (2014) Contrasting primary successional trajectories of fungi and bacteria in retreating glacier soils. *Mol Ecol* **23** : 481–497.
- Buffelskloof Nature Reserve (2019) Buffelskloof private nature reserve.
- Burnham, K.P. and Anderson, D.R. (2002) Model selection and multimodel inference: a practical information-theoretic approach, 2nd ed. New York: Springer-Verlag.
- Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K., et al. (2010) QIIME allows analysis of high-throughput community sequencing data. *Nat Methods* **7** : 335–336.
- Carney, K.M. and Matson, P.A. (2006) The influence of tropical plant diversity and composition on soil microbial communities. *Microb Ecol* **52** : 226–238.
- Chang, C. and HilleRisLambers, J. (2016) Integrating succession and community assembly perspectives. *F1000Research* **5** : F1000 Faculty Rev-2294.
- Chang, C.C. and Turner, B.L. (2019) Ecological succession in a changing world. *J Ecol* **107** : 503–509.
- Christian, N., Sullivan, C., Visser, N.D., and Clay, K. (2016) Plant host and geographic location drive endophyte community composition in the face of perturbation. *Microb Ecol* **72** : 621–632.
- Compant, S., Van Der Heijden, M.G.A., and Sessitsch, A. (2010) Climate change effects on beneficial plant-microorganism interactions. *FEMS Microbiol Ecol* **73** : 197–214.
- Compant, S., Samad, A., Faist, H., and Sessitsch, A. (2019) A review on the plant microbiome: Ecology, functions, and emerging trends in microbial application. *J Adv Res* **19** : 29–37.
- Cordier, T., Robin, C., Capdevielle, X., Desprez-Loustau, M.L., and Vacher, C. (2012) Spatial variability of phyllosphere fungal assemblages: genetic distance predominates over geographic distance in a European beech stand (*Fagus sylvatica*). *Fungal Ecol* **5** : 509–520.
- Cordovez, V., Dini-Andreote, F., Carrión, V.J., and Raaijmakers, J.M. (2019) Ecology and evolution of plant microbiomes. *Annu Rev Microbiol* **73** : 69–88.
- Crowther, T.W., Maynard, D.S., Crowther, T.R., Peccia, J., Smith, J.R., and Bradford, M.A. (2014) Untangling the fungal niche: the trait-based approach. *Front Microbiol* **5** : 1–12.
- Darlison, J., Mogren, L., Rosberg, A.K., Grudén, M., Minet, A., Liné, C., et al. (2019) Leaf mineral content govern microbial community structure in the phyllosphere of spinach (*Spinacia oleracea*) and rocket (*Diplotaxis tenuifolia*). *Sci Total Environ* **675** : 501–512.
- Davey, M., Blaalid, R., Vik, U., Carlsen, T., Kauserud, H., and Eidesen, P.B. (2015) Primary succession of *Bistorta vivipara* (L.) Delabre (Polygonaceae) root-associated fungi mirrors plant succession in two glacial chronosequences. *Environ Microbiol* **17** : 2777–2790.

- David, A.S., Seabloom, E.W., and May, G. (2016) Plant host species and geographic distance affect the structure of aboveground fungal symbiont communities, and environmental filtering affects belowground communities in a coastal dune ecosystem. *Microb Ecol* **71** : 912–926.
- Dini-Andreote, F., Pylro, V.S., Baldrian, P., Van Elsas, J.D., and Salles, J.F. (2016) Ecological succession reveals potential signatures of marine-terrestrial transition in salt marsh fungal communities. *ISME J* **10** : 1984–1997.
- Dini-Andreote, F., Stegen, James, C., van Elsas, Jan, D., and Salles, Joana, F. (2015) Disentangling mechanisms that mediate the balance between stochastic and deterministic processes in microbial succession. *Proc Natl Acad Sci U S A* **112** : E1326–E1332.
- Dong, K., Tripathi, B., Moroenyane, I., Kim, W., Li, N., Chu, H., and Adams, J. (2016) Soil fungal community development in a high Arctic glacier foreland follows a directional replacement model, with a mid-successional diversity maximum. *Sci Rep* **6** : 1–9.
- Elston, D.A., Moss, R., Boulinier, T., Arrowsmith, C., and Lambin, X. (2001) Analysis of aggregation, a worked example: Numbers of ticks on red grouse chicks. *Parasitology* **122** : 563–569.
- Ferrier, S., Manion, G., Elith, J., and Richardson, K. (2007) Using generalized dissimilarity modelling to analyse and predict patterns of beta diversity in regional biodiversity assessment. *Divers Distrib* **13** : 252–264.
- Fitzpatrick, M.C., Mokany, K., Manion, G., Lisk, M., Ferrier, S., and Neito-Lugilde, D. (2021) gdm: Generalized Dissimilarity Modeling. R package version 1.4.2.2.
- Fitzpatrick, M.C., Sanders, N.J., Normand, S., Svenning, J.-C., Ferrier, S., Gove, A.D., and Dunn, R.R. (2013) Environmental and historical imprints on beta diversity: insights from variation in rates of species turnover along gradients. *Proc R Soc B Biol Sci* **280** .
- Fox, J. (2003) Effect displays in R for generalised linear models. *J Stat Softw* **8** : 1–27.
- Fox, J. and Weisberg, S. (2019) An {R} Companion to Applied Regression, Third. Thousand Oaks, California: Sage.
- Gao, C., Montoya, L., Xu, L., Madera, M., Hollingsworth, J., Purdom, E., et al. (2019) Strong succession in arbuscular mycorrhizal fungal communities. *ISME J* **13** : 214–226.
- Gardes, M. and Bruns, T.D. (1993) ITS primers with enhanced specificity for basidiomycetes - application to the identification of mycorrhizae and rusts. *Mol Ecol* **2** : 113–118.
- Glassman, S.I., Wang, I.J., and Bruns, T.D. (2017) Environmental filtering by pH and soil nutrients drives community assembly in fungi at fine spatial scales. *Mol Ecol* **26** : 6960–6973.
- Gotelli, N.J. (2000) Null model analysis of species co-occurrence patterns. *Ecology* **81** : 2606–2621.
- Gotelli, N.J. and Graves, G.R. (1996) Co-occurrence. In *Null Models in Ecology* . Washington: Smithsonian Institution Press, pp. 153–205.
- Greve, M., Gaston, K.J., van Rensburg, B.J., and Chown, S.L. (2008) Environmental factors, regional body size distributions and spatial variation in body size of local avian assemblages. *Glob Ecol Biogeogr* **17** : 514–523.
- Griffin, E. and Carson, W.P. (2018) Tropical tree endophytes: cryptic drivers of forest diversity, species composition, and ecosystem function., pp. 1–52.
- Griffin, E.A., Harrison, J.G., McCormick, M.K., Burghardt, K.T., and Parker, J.D. (2019) Tree diversity reduces fungal endophyte richness and diversity in a large-scale temperate forest experiment. *Diversity* **11** : 1–20.

- Harrison, J.G. and Griffin, E.A. (2020) The diversity and distribution of endophytes across biomes, plant phylogeny and host tissues: how far have we come and where do we go from here? *Environ Microbiol***22** : 2107–2123.
- Hausmann, N.T. and Hawkes, C. V. (2009) Plant neighborhood control of arbuscular mycorrhizal community composition. *New Phytol***183** : 1188–1200.
- Hoch, J.M.K., Rhodes, M.E., Shek, K.L., Dinwiddie, D., Hiebert, T.C., Gill, A.S., et al. (2019) Soil microbial assemblages are linked to plant community composition and contribute to ecosystem services on urban green roofs. *Front Ecol Evol* **7** .
- Hu, Y., Veresoglou, S.D., Tedersoo, L., Xu, T., Ge, T., Liu, L., et al. (2019) Contrasting latitudinal diversity and co-occurrence patterns of soil fungi and plants in forest ecosystems. *Soil Biol Biochem***131** : 100–110.
- Hubbell, S.P. (2001) Unified neutral theory of biodiversity, 1st ed. Princeton: (Princeton University Press.
- Jamison-Daniels, S.-L., Kissling, W.D., Botha, M., Harris, M.A., Gordon, C.E., and Greve, M. (2021) The role of deterministic succession during forest development within a southern African savanna. *Biotropica***53** : 466–476.
- Johnson, A.E. (1979) Succession an unfinished revolution. *Ecology***60** : 240–241.
- Kembel, S.W. and Mueller, R.C. (2014) Plant traits and taxonomy drive host associations in tropical Phyllosphere fungal communities. *Botany* **92** : 303–311.
- Koide, R.T., Ricks, K.D., and Davis, E.R. (2017) Climate and dispersal influence the structure of leaf fungal endophyte communities of *Quercus gambelii* in the eastern Great Basin, USA. *Fungal Ecol***30** : 19–28.
- Laforest-Lapointe, I., Paquette, A., Messier, C., and Kembel, S.W. (2017) Leaf bacterial diversity mediates plant diversity and ecosystem function relationships. *Nature* **546** : 145–147.
- Lau, M.K., Arnold, A.E., and Johnson, N.C. (2013) Factors influencing communities of foliar fungal endophytes in riparian woody plants. *Fungal Ecol* **6** : 365–378.
- Lawson, A., Biggeri, A., Bohnng, D., Lasaffre, E., Viel, J.-E., and Bertollini, R. (eds). (1999) Disease mapping and risk assessment for public health, 1st ed. Chichester, UK: John Wiley.
- Legendre, P. and Borcard, D. (2008) Analyzing or Explaining Beta Diversity? Comment. *Ecology* **89** : 3238–3244.
- Legendre, P. and Legendre, L. (2012) Numerical Ecology, 3rd ed. Elsevier.
- Liu, J., Zhao, J., Wang, G., and Chen, J. (2019) Host identity and phylogeny shape the foliar endophytic fungal assemblages of *Ficus*. *Ecol Evol* **9** : 10472–10482.
- McCulloch, C.E. (1997) An Introduction to Generalized Linear Mixed Models.
- McMurdie, P.J. and Holmes, S. (2014) Waste Not, Want Not: Why Rarefying Microbiome Data Is Inadmissible. *PLOS Comput Biol* **10** : e1003531.
- Morisita, M. (1962) I²-Index, a measure of dispersion of individuals. *Popul Ecol* **4** : 1–7.
- Mucina, L. and Rutherford, M.C. (2011) The vegetation of South Africa, Lesotho and Swaziland., Reprint. Pretoria: Strelitzia 19. South African National Biodiversity Institute.
- Nakagawa, S., Johnson, P.D.C., and Schielzeth, H. (2017) The coefficient of determination R² and intra-class correlation coefficient from generalized linear-mixed effects models revisited and expanded. *J R Soc Interface* **14** .
- Nekola, J.C. and White, P.S. (1999) The distance decay of similarity in biogeography and ecology. *J Biogeogr* **26** : 867–878.

- Nemergut, D.R., Schmidt, S.K., Fukami, T., O'Neill, S.P., Bilinski, T.M., Stanish, L.F., et al. (2013) Patterns and Processes of Microbial Community Assembly. *Microbiol Mol Biol Rev* **77** : 342–356.
- Oksanen, J.F., Blanchet, G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., et al. (2019) vegan: Community Ecology Package.
- Oono, R., Lefevre, E., Simha, A., and Lutzoni, F. (2015) A comparison of the community diversity of foliar fungal endophytes between seedling and adult loblolly pines (*Pinus taeda*). *Fungal Biol* **119** : 917–928.
- Oono, R., Rasmussen, A., and Lefevre, E. (2017) Distance decay relationships in foliar fungal endophytes are driven by rare taxa. *Environ Microbiol* **19** : 2794–2805.
- Peay, K.G., Kennedy, P.G., and Talbot, J.M. (2016) Dimensions of biodiversity in the Earth mycobiome. *Nat Rev Microbiol* **14** : 434–447.
- Polme, S., Bahram, M., Jacquemyn, H., Kennedy, P., Kohout, P., Moora, M., et al. (2018) Host preference and network properties in biotrophic plant–fungal associations. *New Phytol* **217** : 1230–1239.
- R Core Team (2019) R: A language and environment for statistical computing.
- Raup, D.M. and Crick, R.E. (1979) Measurement of faunal similarity in paleontology. *J Paleontol* **53** : 1213–1227.
- Reich, P.B., Ellsworth, D.S., Walters, M.B., Vose, J.M., Gresham, C., Volin, J.C., and Bowman, W.D. (1999) Generality of leaf trait relationships: A test across six biomes. *Ecology* **80** : 1955–1969.
- Salles, J.F., Poly, F., Schmid, B., and Roux, X. Le (2009) Community niche predicts the functioning of denitrifying bacterial assemblages. *Ecology* **90** : 3324–3332.
- Santoyo, G., Hernandez-Pacheco, C., Hernandez-Salmeron, J., and Hernandez-Leon, R. (2017) The role of abiotic factors modulating the plant-microbe-soil interactions: Toward sustainable agriculture. A review. *Spanish J Agric Res* **15** : 1–15.
- Schneider, C.A., Rasband, W.S., and Eliceiri, K.W. (2012) NIH Image to ImageJ: 25 years of image analysis. *Nat Methods* **9** : 671–675.
- Scholtysik, A., Unterseher, M., Otto, P., and Wirth, C. (2013) Spatio-temporal dynamics of endophyte diversity in the canopy of European ash (*Fraxinus excelsior*). *Mycol Prog* **12** : 291–304.
- Soininen, J., McDonald, R., and Hillebrand, H. (2007) The distance decay of similarity in ecological communities. *Ecography (Cop)* **30** : 3–12.
- Tellez, P.H., Woods, C.L., Formel, S., and Van Bael, S.A. (2020) Relationships between foliar fungal endophyte communities and ecophysiological traits of CAM and C3 epiphytic bromeliads in a neotropical rainforest. *Diversity* **12** : 1–15.
- Terhonen, E., Blumenstein, K., Kovalchuk, A., and Asiegbu, F.O. (2019) Forest tree microbiomes and associated fungal endophytes: Functional roles and impact on forest health. *Forests* **10** : 1–32.
- Tripathi, B.M., Stegen, J.C., Kim, M., Dong, K., Adams, J.M., and Lee, Y.K. (2018) Soil pH mediates the balance between stochastic and deterministic assembly of bacteria. *ISME J* **12** : 1072–1083.
- Turner, B.L., Zemunik, G., Laliberte, E., Drake, J.J., and Jones, F.A. (2019) Contrasting patterns of plant and microbial diversity during long-term ecosystem development. *J Ecol* **107** : 606–621.
- U'Ren, J.M., Lutzoni, F., Miadlikowska, J., Zimmerman, N.B., Carbone, I., May, G., and Arnold, A.E. (2019) Host availability drives distributions of fungal endophytes in the imperilled boreal realm. *Nat Ecol Evol* **3** : 1430–1437.

- Unterseher, M., Peršoh, D., and Schnittler, M. (2013) Leaf-inhabiting endophytic fungi of European Beech (*Fagus sylvatica* L.) co-occur in leaf litter but are rare on decaying wood of the same host. *Fungal Divers* **60** : 43–54.
- Unterseher, M., Petzold, A., and Schnittler, M. (2012) Xerotolerant foliar endophytic fungi of *Populus euphratica* from the Tarim River basin, central China are conspecific to endophytic ITS phylotypes of *Populus tremula* from temperate Europe. *Fungal Divers* **54** : 133–142.
- Unterseher, M., Reiher, A., Finstermeier, K., Otto, P., and Morawetz, W. (2007) Species richness and distribution patterns of leaf-inhabiting endophytic fungi in a temperate forest canopy. *Mycol Prog* **6** : 201–212.
- Unterseher, M., Siddique, A.B., Brachmann, A., and Peršoh, D. (2016) Diversity and composition of the leaf mycobiome of beech (*Fagus sylvatica*) are affected by local habitat conditions and leaf biochemistry. *PLoS One* **11** .
- Vaz, A.B.M., Fontenla, S., Rocha, F.S., Brandão, L.R., Vieira, M.L.A., de Garcia, V., et al. (2014) Fungal endophyte β -diversity associated with Myrtaceae species in an Andean Patagonian forest (Argentina) and an Atlantic forest (Brazil). *Fungal Ecol* **8** : 28–36.
- Vellend, M. (2010) Conceptual synthesis in community ecology. *Q Rev Biol* **85** : 183–206.
- Venables, W.N. and Ripley, B.D. (2002) Modern Applied Statistics with S, 4th ed. Springer Berlin Heidelberg.
- Veteli, T.O., Kuokkanen, K., Julkunen-Tiitto, R., Roininen, H., and Tahvanainen, J. (2002) Effects of elevated CO₂ and temperature on plant growth and herbivore defensive chemistry. *Glob Chang Biol* **8** : 1240–1252.
- Vincent, J.B., Weiblen, G.D., and May, G. (2016) Host associations and beta diversity of fungal endophyte communities in New Guinea rainforest trees. *Mol Ecol* **25** : 825–841.
- White, T.J., Bruns, T.D., Lee, S., and Taylor, J.W. (1990) Amplification and direct sequencing of fungal ribosomal RNA genes for phylogenetics. In *PCR Protocols: A Guide to Methods and Applications* . Innis, M., Gelfand, D., Shinsky, J., and White, T.J. (eds). San Diego, pp. 315–322.
- Williams, K.J., Belbin, L., Austin, M.P., Stein, J.L., and Ferrier, S. (2012) Which environmental variables should I use in my biodiversity model? *Int J Geogr Inf Sci* **26** : 2009–2047.
- Willis, A.D. (2019) Rarefaction, Alpha Diversity, and Statistics. *Front Microbiol* **0** : 2407.
- Yao, H., Sun, X., He, C., Maitra, P., Li, X.-C., and Guo, L.-D. (2019) Phyllosphere epiphytic and endophytic fungal community and network structures differ in a tropical mangrove ecosystem. *Microbiome* **7** : 57.
- Zhou, J. and Ning, D. (2017) Stochastic community assembly: does it matter in microbial ecology? *Microbiol Mol Biol Rev* **81** : pii: e00002-17.
- Zimmerman, N.B. and Vitousek, P.M. (2012) Fungal endophyte communities reflect environmental structuring across a Hawaiian landscape. *Proc Natl Acad Sci* **109** : 13022–13027.
- Zuur, A., Ieno, E.N., Walker, N., Saveliev, A.A., and Smith, G.M. (2009) Mixed effects models and extensions in ecology with R, 1st ed. New York: Springer.

Figure legends

Figure 1. Effect plot showing the relationship between Raup-Crick similarity values of foliar fungal endophyte communities on *Euclea crispa* and BC area as a proxy for age of the BCs. Light blue shading represents the 95% confidence interval for the fitted model. Only significant trends are shown.

Figure 2. a-f) Significant retained predictor variables for the GDM performed on the full dataset. Each panel (a-f) represents the fitted I-splines (partial regression fits) of each retained predictor variable associated with the compositional turnover in foliar fungal endophyte communities. The maximum height reached by each curve represents the total amount of compositional turnover explained by that variable, holding all other variables constant.

Figure 3. a-c) Significant retained predictor variables for the GDM performed on the endophyte communities within the host *Euclea crispa*. d & e) Significant retained predictor variables for the GDM performed on the endophyte communities within the host *Canthium inerme*. f) Significant retained predictor variable for the GDM performed on the endophyte communities with the host *Searsia chirindensis*. Each panel (a-f) represents the fitted I-splines (partial regression fits) of each retained predictor variable associated with the compositional turnover in foliar fungal endophyte communities. The maximum height reached by each curve represents the total amount of compositional turnover explained by that variable, holding all other variables constant.

Figure 4. Predictor effect plots for the best subset GLMM model for ASV richness of foliar fungi. a) ASV richness decreased with light intensity - the x-axis represents inverse light intensity measured as photosynthetic active radiation (PAR) outside a BC minus PAR inside the BC. b) Maximum temperatures experienced within the BCs had no effect on ASV richness in *C. inerme* and *E. crispa*, but decreased with maximum temperature in *S. chirindensis*. c) Height of the trees from which samples were taken had no effect on ASV richness in *C. inerme* and *E. crispa*, whereas ASV richness decreased with maximum temperature in *S. chirindensis*. Light blue shading represents the 95% confidence interval for the fitted model.





