

Sediment oxygen uptake and hypoxia: a simple mass-balance model for estuaries and coastal oceans

Jing Sun^{1,2}, Liuqian Yu^{2,3*}, Xingyu Yang^{1,2}, Jianping Gan^{1,2}, Hongbin Yin⁴, and Jiying Li^{1,2*}

¹Department of Ocean Science, The Hong Kong University of Science and Technology, Clear Water Bay, Kowloon, Hong Kong SAC, P. R. China

²Center for Ocean Research in Hong Kong and Macau, Hong Kong SAC, P. R. China

³Earth, Ocean and Atmospheric Sciences Thrust, The Hong Kong University of Science and Technology (Guangzhou), Guangdong, P.R. China

⁴Nanjing Institute of Geography and Limnology, Chinese Academy of Science, Nanjing, Jiangsu, P. R. China

Corresponding author:

Jiying Li (jiyingli@ust.hk; ORCID: [0000-0003-1677-6922](https://orcid.org/0000-0003-1677-6922))

Liuqian Yu (liuqianyu@hkust-gz.edu.cn; ORCID: [0000-0002-5492-8213](https://orcid.org/0000-0002-5492-8213))

Key Points:

- Sediment respire >60% of the organic matter produced in the water column of the Pearl River Estuary (PRE) region, leading to high sediment oxygen uptake.
- The sediment's effect on the bottom oxygen loss is controlled by the thickness of the bottom boundary layer.
- We develop a simple and generic mass-balance model to understand hypoxia conditions and timescales in the PRE and similar coastal systems.

Abstract

Hypoxia is increasing in coastal oceans. This is because eutrophication has increased oxygen consumption, while less oxygen is replenished to the bottom under stronger stratification. Quantifying these biogeochemical and physical drivers is important for management and predicting future trends. By using observations from the Pearl River Estuary (PRE) region (10-70 m deep) and similar coastal systems, this paper introduces a simple analysis to quantify both the biogeochemical and physical drivers of hypoxia. We show that in the PRE region, sediment respire >60% of organic matter produced in the water column, leading to high sediment oxygen uptake (average $41.1 \pm 16.3 \text{ mmol m}^{-2} \text{ d}^{-1}$) and shallow oxygen penetrations (2-7 mm). The sediment's effect on the bottom oxygen loss becomes stronger with the reducing thickness of the bottom boundary layer. We then construct a generic mass-balance model to quantify oxygen loss, determine timescales of hypoxia formation, and explain within- and cross-system variabilities.

Plain language summary

Coastal oceans, especially those off rivers and estuaries, frequently experience low-oxygen conditions such as hypoxia (dissolved $\text{O}_2 < 2 \text{ mg L}^{-1}$). This is because the nutrient-rich coastal ocean produces high amount of organic matter, which settles to the bottom waters to consume oxygen. Hypoxia becomes more severe if the resupply of oxygen from the atmosphere and the surface water to the bottom water is blocked by strong density stratification. To quantify these effects, we study a typical coastal system, the Pearl River Estuary and the adjacent shelf, combining field and lab observations and mass-balance modeling. We find that sediment consumes a substantial amount of oxygen, and its contribution to hypoxia can be predicted if we know how thick the stratified bottom layer is. We derive some simple equations to understand oxygen loss, which can tell us what level of oxygen consumption can render the system hypoxia, and for how long the stratification needs to be maintained for hypoxia to develop. We also show that the model is generic and can be applied to other similar coastal systems, such as the northern Gulf of Mexico and the Changjiang Estuary region, to explain the variability in hypoxia conditions and timescales.

1 Introduction

Low oxygen conditions are becoming more frequent in coastal oceans, affecting biogeochemical cycles and marine life (Katsev et al., 2007; Levin et al., 2009; Middelburg & Levin, 2009). These conditions are a result of high oxygen consumption and weak ventilation, especially in bottom waters where oxygen depletion can be severe, leading to hypoxia ($O_2 < 2 \text{ mg L}^{-1}$). In estuaries and coastal shelves, oxygen is rapidly consumed due to the high organic matter production driven by terrestrial and/or upwelling nutrients; freshwater inputs also enhance vertical density stratification and reduce the oxygen resupply to the bottom. The increasing anthropogenic fertilization of the coastal ocean, together with the stronger stratification caused by global warming, therefore, can intensify hypoxia in both magnitude and duration (Diaz & Rosenberg, 2008).

Quantifying these physical and biogeochemical drivers of hypoxia is important for management and predicting future trends. In addition to monitoring oxygen (Grégoire et al., 2021), mechanistic modeling becomes very useful (Laurent & Fennel, 2017; Peña et al., 2010). These models simulate the oxygen budget considering its source/sink components, including physical transports, air-sea exchanges, photosynthesis, and biogeochemical consumption in the water column and sediments (Peña et al., 2010; Testa et al., 2017). Among models of various complexity, coupled-physical-biogeochemical models are the most comprehensive, for they simulate the interactions between nutrient-driven organic matter productivity and physical transport, the two major drivers of oxygen variability (Fennel et al., 2016; Yu et al., 2015a, 2021). These models, however, rely on parametrizations of many processes, for example, the relationships between lights, temperature, nutrients, organic matter production, respiration, food-web dynamics, and the reactions in the sediments and benthic-pelagic exchanges (Peña et al., 2010; Yu et al., 2015a, b, 2021). Moreover, parametrizations can be system-specific and need to be well constrained by a large matrix of observations (e.g., dynamics in biomass indicators, nutrients, and physical properties). Thus, the models usually have low transferability among systems (Peña et al., 2010). For cross-system comparison, simple scale analyses are useful (Fennel & Testa, 2019). For example, the vulnerability of the system can be understood by comparing the timescale of hypoxia formation to the water residence time of the region (Fennel & Testa, 2019). However, while such an approach provides regional order-of-magnitude understanding, it is inadequate for describing local heterogeneity, which is important for estuarine coastal systems.

This paper introduces a simple analysis that can describe regional heterogeneity and also enables cross-system comparison. By using water column and sediment oxygen data across a typical estuary and coastal shelf, we derive a simple mass-balance model to understand oxygen loss without a detailed formulation of the physical and biogeochemical complexity. The model can be used to understand the sensitivity of the system to both the biogeochemical and physical drivers of hypoxia. We then discuss how the model can be generalized to understand within-and cross-system variability,

using examples from similar coastal systems including the Northern Gulf of Mexico, the Changjiang Estuary region, and the Chesapeake Bay.

2. Field Methods

We study the Pearl River Estuary and adjacent shelf waters (PRE region) in the summer of 2021 (Fig.1a and Table.S1; Li et al. 2024). We used a SeaBird-SBE17-plus conductivity–temperature–depth (CTD) to measure the water column temperature, density, salinity, O₂ concentrations and calculated the buoyancy frequency (N^2 ; s⁻²):

$$N^2 = \frac{g}{\rho(z)} \frac{d\rho(z)}{dz} \quad \text{Eq.1}$$

g is gravity acceleration; z is water depth; ρ is potential density. High N^2 means high stability of the water column.

Sediment cores with undisturbed overlying waters were collected using a Uwitec corer. Oxygen micro-profiles were obtained using a Unisense O₂ electrode. The sediment-water interface has the sharpest oxygen gradient. Oxygen penetration depth was defined as the depth where oxygen is under the detection limit of ~0.3 μmol L⁻¹. Total sediment oxygen uptake (SOU; mmol m⁻² d⁻¹), defined as the downward flux of oxygen into the sediment, was determined using onboard whole-core incubations: sediment cores were stabilized, sealed, and monitored for the oxygen concentrations in the overlying waters, which were gently stirred to generate the water movement and create a diffusive boundary layer (Bowman & Delfino, 1980; Glud, 2008). SOU was calculated from the linear decrease of oxygen. The incubation typically lasts 2-3 hours. Detailed methods are described in SI.1.

3. Results

3.1 Stratification and hypoxia in the water column – The water column develops hypoxia within the nearshore waters (10-20 m) off the estuary during the summer (Fig.1a), whereas offshore sites (>30 m) are well oxygenated. Such local seasonal hypoxia has been observed for several recent years (Li et al., 2020; Yu et al., 2021). At most sites, the water column is stratified by vertical gradients of temperature and salinity (Figs.1b and S1): the surface water from the upstream is warmer and fresher, while the bottom seawater is colder and saltier. In the stratified waters, the surface mixed layers are thin, with the density drastically increasing below 5-10 m forming a pycnocline (Fig.S1). This parallels the changes of buoyancy frequency (N^2) (Fig.1c), which peaks within the pycnocline and decreases downward, until the density has little variation, forming a stable bottom boundary layer (BBL; Figs.1c and S1) (Trowbridge & Lentz, 2018). We defined the upper boundary of BBL as the depth where N^2 increases drastically above (i.e., $N^2 > 0.01$ s⁻²). The BBL has homogenously lower oxygen and reaches hypoxia at some sites (Figs.1a and 1c).

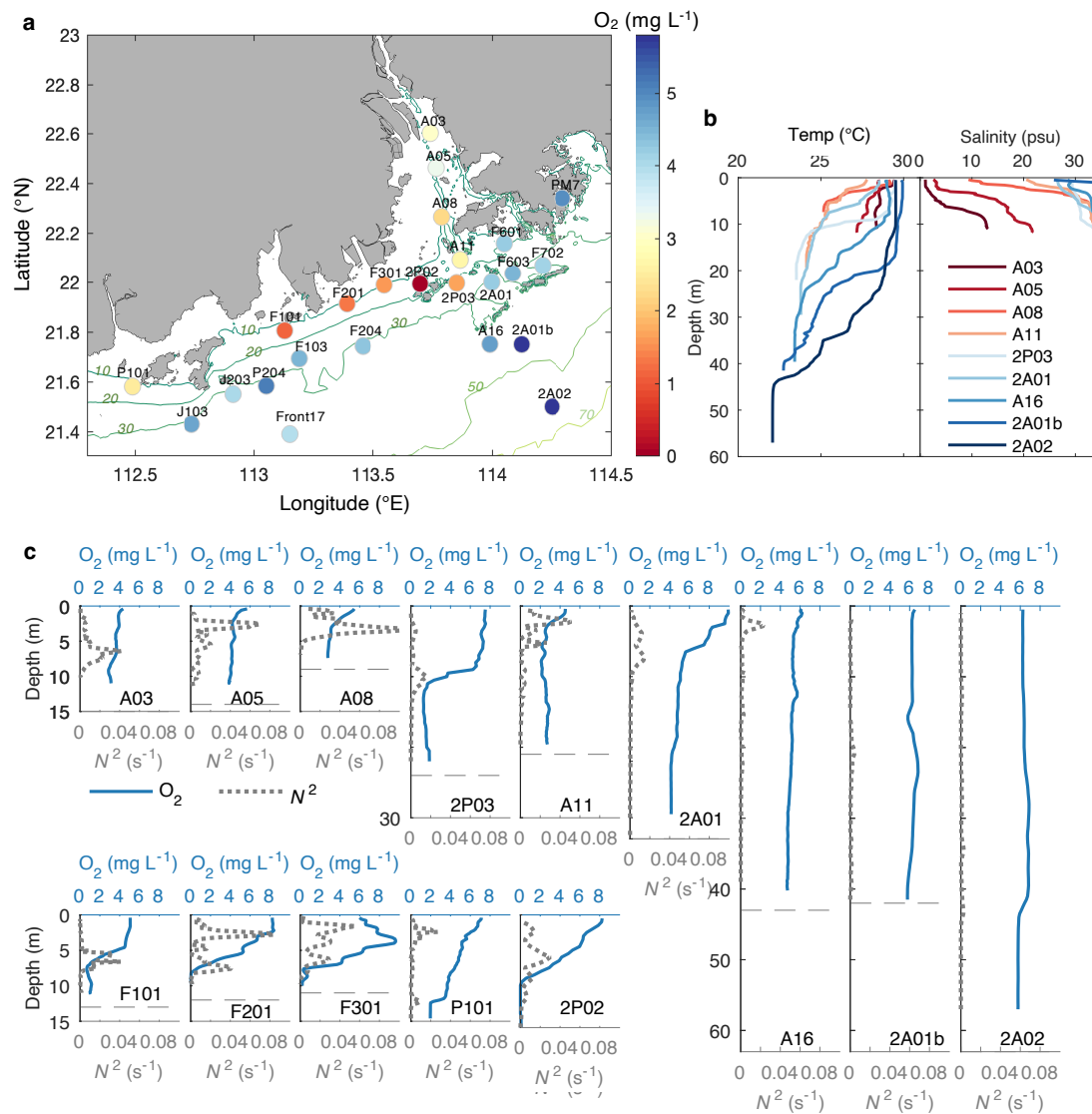


Figure 1 Water column physiochemistry in the Pearl River Estuary region. a) Sampling locations and bottom oxygen concentrations; b) Temperature and salinity profiles across the estuary to open ocean gradient c) Buoyancy frequency (N^2) and O_2 profiles across the salinity gradient and at the hypoxic sites. Horizontal dashed lines represent the bottom.

3.2 Sediment oxygen uptake and bottom water oxygen – Oxygen concentration in the BBL is controlled by the flux from the upper layer, which can be restricted by stratification, and the consumption in the water and sediments. In the sediments, oxygen drops sharply from 75-150 $\mu\text{mol L}^{-1}$ near the interface to nondetectable ($<0.3 \mu\text{mol L}^{-1}$) within a few millimeters downcore (1-7 mm; Figs.2a and S3). Oxygen penetration decreases with decreasing water depths, consistent with global observations but more dramatically compared to the open ocean (Figs. S4b and S4c). The drastic depletion of oxygen in sediments is a result of high SOU ($16.5\text{-}70.5 \text{ mmol m}^{-2} \text{ d}^{-1}$, average 41.1 ± 16.3

mmol m⁻² d⁻¹; Fig.2b and Table.S1), consistent with typical coastal sediments (Fig.S4d). Nearshore sediments have relatively higher SOU compared to offshore (Fig.2b). If we assume oxygen is predominantly consumed via organic matter respiration, this consumption averages to ~62±25% of the organic matter produced in the water column of the region (66 mmol m⁻² d⁻¹) (Cai et al., 2004).

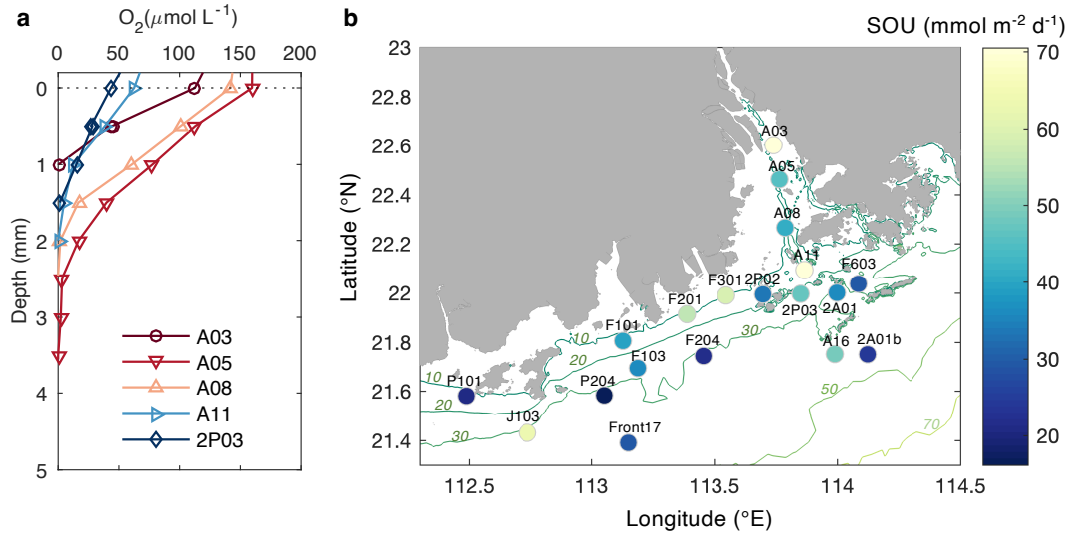


Figure 2 Sediment oxygen distribution and uptake. a) O₂ profiles at selected sites; **b)** sediment oxygen uptake (SOU).

High SOU would contribute significantly to the oxygen loss in the bottom water. However, the correlation between SOU and bottom oxygen is weak (Fig.S5 and Yu et al. (2015a)). This is because the thickness of the BBL would affect the sediment's impact: a thin layer would respond faster to SOU, while a thicker bottom reacts less with the effect of SOU being diluted. Indeed, the hypoxia sites have thin BBL (1.3 to <10 m), whereas the well-oxygenated sites have thicker BBL or entirely mixed water column (Figs.1c, S2, and Table.S1). By normalizing SOU to the thickness of the BBL (h), we obtain the oxygen depletion rate in the BBL by the sediments (SOU/ h ; mmol m⁻³ d⁻¹), which exhibits a strong relationship with the bottom O₂ level (Fig.3a). A stronger correlation appears when considering the integrated O₂ over the BBL (mmol m⁻²): oxygen decreases exponentially with increasing SOU/ h (Fig.3b). The relationships exist in similar systems such as the northern Gulf of Mexico and the Changjiang Estuary region (Figs.3a and 3b); both are estuary dominated shelves experiencing hypoxia (McCarthy et al., 2013; Zhu et al. 2016).). Similar relationships are also seen for the entire water column (Fig.S6).

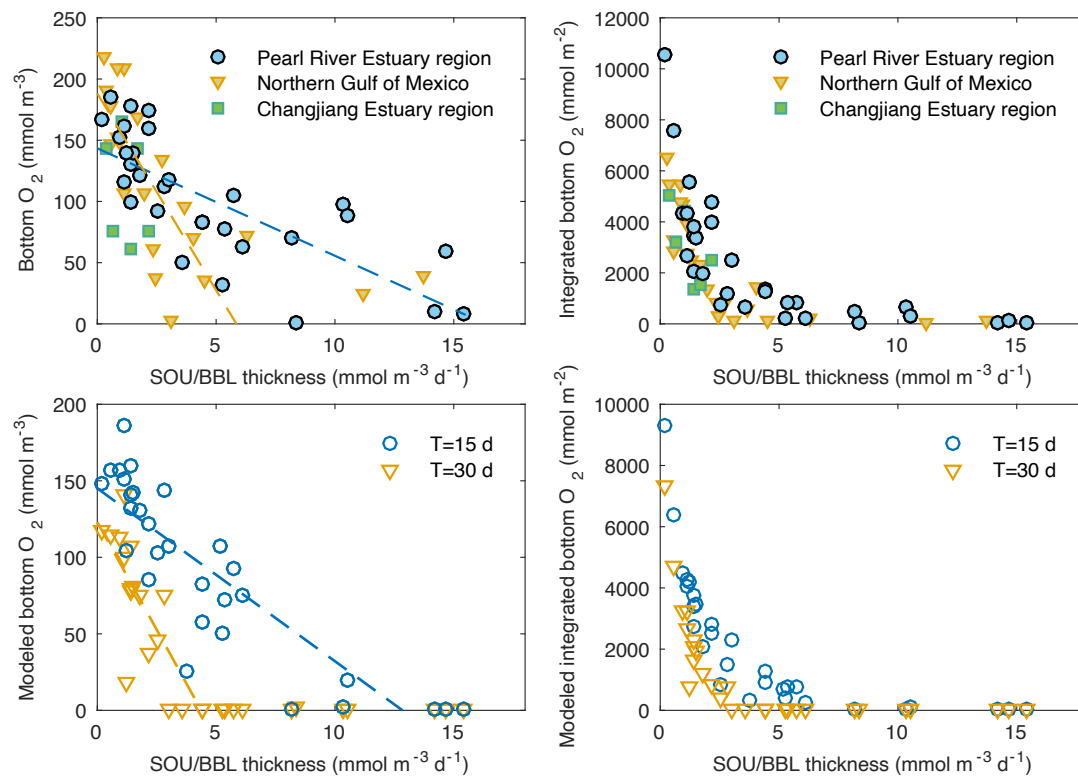


Figure 3 Bottom oxygen vs SOU normalized to BBL thickness (SOU/h). a and b) Bottom oxygen vs SOU/h in the PRE region, North Gulf of Mexico (McCarthy et al., 2013), and Changjiang Estuary region (Zhang et al., 2017); c and d) Modeled bottom oxygen vs SOU/h, using water stability periods of T= 15 and 30 days.

The water column oxygen uptake (WOU) is not explicitly considered, but it does not mean WOU is not important and omitted. Instead, the robust relationships suggest an intrinsic connection between SOU and WOU, which are both driven by the degradation of organic matter. Their proportions are determined by the water thickness, which determines the time settling particles spend in the water and thus their leftovers to drive SOU. This explains the curvature in Fig.3b: at the sites with low SOU/h, the BBL is thick and thus WOU becomes important, making the oxygen level sensitive to the apparent per-SOU change (steeper slopes at low SOU/h in Figs.3b).

The strong correlations between oxygen level and SOU/h also suggest the possibility of parameterizing SOU. SOU is important in controlling the water column oxygen budget and sediment geochemistry and fluxes and rates (e.g., denitrification, sulfate reduction) (Li et al., 2018a, b; Li & Katsev, 2014), but published SOC data is scarce due to measurement challenges. The potential parameterization of SOU using more obtainable water column CTD data is useful.

4. The Mass-Balance Model — Inspired by the observations above, we construct a mass-balance model to explore their physical meanings and quantitative insights. Details

of the model are narrated in SI.2, which we briefly introduce here to facilitate discussion. The change of O₂ concentration in the BBL is caused by fluxes from above (assumed to be small and neglected here due to the large N₂ above the BBL) and the sediment (F_{s-O_2} , which equals -SOU), and the reaction within the BBL (F_{BBL-O_2} , which equals -WOU_{BBL}):

$$\frac{d\bar{C}_{BBL-O_2}}{dt} = \frac{F_{s-O_2}}{h} + \frac{F_{BBL-O_2}}{h} \quad \text{Eq.2}$$

Here, \bar{C}_{BBL-O_2} is the average O₂ concentration in the BBL. Assuming oxygen is mostly consumed to respire organic matter (Zhang & Li, 2010), F_{BBL-O_2} depends on the amount of organic carbon settling into the BBL (C_{w-C}^{H-h} ; H is the total depth and h is the thickness of the BBL, hence the upper boundary of BBL is at $H-h$; the subscript 'w-C' stands for water-column carbon), the reaction rate (k_w), and the time the particles spend in the BBL ($t=h/u_w$; u_w is the particle settling velocity). Considering a 1C:1O₂ ratio and first-order reaction for simplicity, the reaction in the BBL is

$$F_{BBL-O_2} = \int_{H-h}^H -k_w C_{w-C} dz_w = \int_{H-h}^H -k_w C_{w-C}^H e^{-\frac{k_w}{u_w}(z_w-H+h)} dz_w = u_w C_C^{H-h} \left(e^{-\frac{k_w}{u_w}h} - 1 \right) = -WOU_{BBL} \quad \text{Eq.3}$$

where C_{w-C} is the concentration of organic matter in the water column; z_w is the vertical displacement. Similarly, for pseudo-steady-state consideration, the O₂ flux into the sediment equals the organic carbon flux ($u_w C_{w-C}^H$; C_{w-C}^H is the concentration at the sediment-water interface at depth H) corrected by a recycling efficiency (ε) for not all organic matter reaching sediments being reactive (Li et al., 2012):

$$F_{s-O_2} = -\varepsilon u_w C_C^H = -SOU \quad \text{Eq.4}$$

Because sedimentation ($u_w C_C^H$) is from the leftover of remineralization in the water column:

$$u_w C_{w-C}^H = u_w C_{w-C}^{H-h} + F_{BBL-O_2} \quad \text{Eq.5}$$

Combining Eqs.3, 4, and 5,

$$F_{s-O_2} = -\varepsilon u_w C_{w-C}^H = -\varepsilon (u_w C_{w-C}^{H-h} + F_{BBL-O_2}) = -\varepsilon u_w C_{w-C}^{H-h} e^{-\left(\frac{k_w}{u_w}\right)h} \quad \text{Eq.6}$$

Therefore, from Eqs. 3 and 6,

$$F_{BBL-O_2} = \frac{1}{\varepsilon} \left(e^{\left(\frac{k_w}{u_w}\right)h} - 1 \right) F_{s-O_2} = \frac{1}{\varepsilon} (e^{fh} - 1) F_{s-O_2} \quad \text{Eq.7.1}$$

$$\text{or } WOU_{BBL} = \frac{1}{\varepsilon} (e^{fh} - 1) SOU \quad \text{Eq.7.2}$$

For simplicity, we introduce a parameter, spatial reactivity ($f = k_w/u_w$, m⁻¹), which describes the reaction proceeded per-unit distance the particles move downwards. Eq.7.2 shows how WOU_{BBL} and SOU are related (see observations above). We then rewrite the oxygen budget by combining Eq.7 and Eq.2 and integrating it over a period (T), during which the present oxygen level develops:

$$\text{AOU}_{\text{BBL}} h = T \left(1 + \frac{1}{\varepsilon} (e^{fh} - 1) \right) \text{SOU} \quad \text{Eq.8}$$

AOU_{BBL} is the apparent oxygen utilization in the BBL. In summary, oxygen utilization in the BBL is determined by the SOU, the spatial reactivity of organic matter (f), the thickness of BBL (h), the recycling efficiency in sediments (ε), and the duration (T) of stratification to develop the AOU. Similarly, for the whole water column,

$$\text{AOU} \times H = T \left(1 + \frac{1}{\varepsilon} (e^{fH} - 1) \right) \text{SOU} \quad \text{Eq.9}$$

$$\text{and } \text{WOU} = \frac{1}{\varepsilon} (e^{fH} - 1) \text{SOU} \quad \text{Eq.10}$$

One may fit the data (SOU, h , H , and AOU) to estimate the parameters (f , ε , and T). However, as the data is limited and scattered, fitting them to a model with multiple parameters will lead to overfitting producing wrong results. As we are interested in the physical insights rather than definite values (also not realistic as they are naturally variable), we choose to fix the sediment recycling efficiency as $\varepsilon=0.5$, for that the deep sediments bury about half of the organic matter reaching the seafloor (Zhou, 2022), consistent with observations in shallow-oxygenated sediments (Li et al., 2018b). We use $T=15$ days for the lack of information, but also constrained by observations that stratification and hypoxia take two weeks to redevelop after the water column is mixed up (Zhao et al., 2021). Under these constraints, we obtain an organic matter spatial reactivity of $f=0.026 \text{ m}^{-1}$ for the PRE region, and the model reproduces the water column AOU well (Fig.S7). The estimate is consistent with the rate of organic matter remineralization estimated for water depth (H) of 10-20 m and that 62% of the organic matter reaches the seafloor: $f = (1-0.62)/H = 0.014\text{-}0.028 \text{ m}^{-1}$ (details in SI.3).

5. Discussion

5.1 Conditions and time scales for hypoxia— The model explains the relationships between the bottom O_2 level and SOU/h in the PRE region and other similar systems (Figs.3c and 3d). The duration of stratification determines the sensitivity of O_2 levels to SOU/h : the northern Gulf of Mexico has long periods of stratification (Bianchi et al., 2010; Dzwonkowski et al., 2018) and thus appears more sensitive to SOU: the bottom oxygen drops more with the same level of increase in SOU/h (Figs.3a and 3b). This is consistent with model results when the stratification period is set to be longer (e.g., $T=30$ versus 15 d; Figs.3c and 3d). Alternatively, the high sensitivity of oxygen can be achieved in systems with higher spatial reactivity of settling organic matter (f) (Fig.S8).

The model provides several quantitative insights. As expected, AOU_{BBL} increases with increasing SOU and the effect is more dramatic when BBL is thin (Fig.4a). In the PRE region, a BBL of <10 m is susceptible to hypoxia even at low SOU (Fig.4a). For thicker BBL to reach hypoxia, higher SOU is required, but the effect is not linear: in the PRE region, when SOU is above $\sim 75 \text{ mmol m}^{-2} \text{ d}^{-1}$, the bottom can easily develop hypoxia regardless of the BBL thickness (Fig.4a). This is because, in environments with

organic matter deposition supporting such high SOU, the WOU would be proportionally high (Eq.7.2).

We can estimate the time required for developing hypoxia (Figs.4b): waters with high SOU and thin BBL become hypoxic within a shorter time. In the PRE region, it requires $< \sim 15$ days for thin BBLs ($< \sim 10$ m) to develop hypoxia; the thicker BBLs require longer, but in general not longer than 60 days (Fig.4b). Assuming similar organic matter spatial reactivity (f) in other coastal waters (but see discussion later for variability), we can estimate their hypoxia time scales (T_{hyp} ; Fig.4c), which can be compared to the stratification time scales to see if hypoxia can develop. For example, in the seasonally hypoxic northern Gulf of Mexico, most hypoxia sites have T_{hyp} of < 30 –40 days, which is generally shorter than the duration of stratification (30 days to several months) (Bianchi et al., 2010). In the Chesapeake Bay mid-stem central channel, the pycnocline remains stable for several summer months (Boynton et al., 2022), thus hypoxia persistently develops even under low SOU (Fig.4c). More extreme cases are Gulf of St. Lawrence, the Black Sea, and the Baltic Sea, where persistent and even permanent stratification sustains thick hypoxia layers under low SOU (Fig.4c).

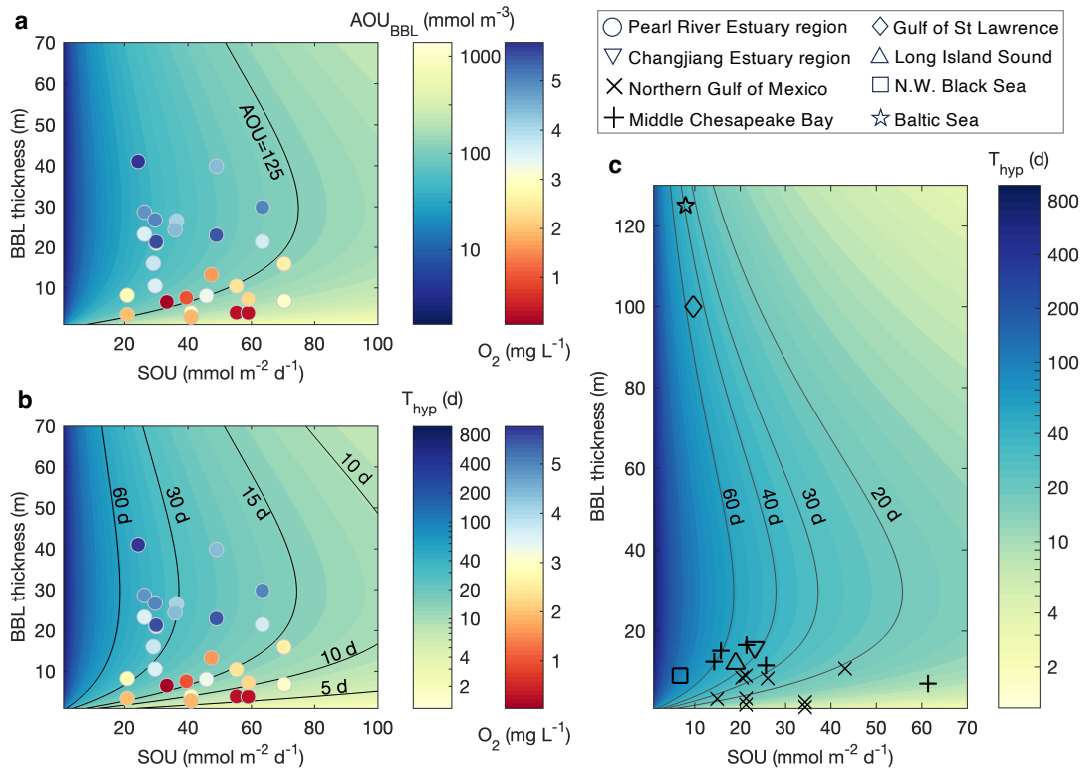


Figure 4 Modeled apparent oxygen utilization (AOU) and time scales for developing hypoxia (T_{hyp}). **a)** AOU_{BBL} as a function of SOU and BBL thickness in the PRE region ($f=0.026 m^{-1}$ and $T=15 d$). Higher AOU indicates higher oxygen loss and hypoxia occurs when $AOU_{BBL} > 125 mmol m^{-3}$ (observed bottom O_2 levels are shown with filled colors for comparison); **b)** the time required for the BBL to become hypoxia (T_{hyp}) ($f=0.026 m^{-1}$). If

stratification sustains longer than T_{hyp} , the bottom water becomes hypoxic. The sites with thick BBL and low SOU need a longer time to develop hypoxia thus the oxygen concentrations remain high. **c)** same as **b** but hypoxia sites from other estuarine and coastal systems (Table S3) (Boynton et al., 2019, 2022; Fennel & Testa, 2019; McCarthy et al., 2013; Zhang et al., 2017).

5.2 Sediments vs water column oxygen uptake— The model can estimate the contribution of SOU to total oxygen loss (%SOU) and explain the global observations in estuaries and coasts (Boynton et al., 2018) (Fig.5): %SOU declines exponentially with the water layer thickness (for whole water column or BBL). Intuitively, the organic matter spends more time in thicker waters consuming oxygen before reaching the sediments, thus reducing the %SOU. The model (Eqs.7.2 and 10) formulates this understanding and explains the possible variabilities: %SOU is regulated by the organic matter spatial reactivity (f) and sediment recycling efficiency (ε). In systems with slowly settling particles (low u_w), the large f (k_w/u_w) would lead to low %SOU (Fig.5). In contrast, SOU contributes more when sedimentation is rapid (large u_w and thus small f ; Fig.5). Particle settling velocity is controlled by particle concentrations (Archer & Devol, 1992), but more importantly by turbulence intensity, which can either accelerates settling by promoting particle collisions and flocculation (Ruiz et al., 2004) or decrease it by breaking up floc when shear stress is too large (Manning, 2004). In rivers and estuaries, strong turbulence also leads to particle resuspension (Boynton et al., 2018), increasing the time the particles spend in the water column to consume oxygen (Moriarty et al. 2021). This might explain why rivers and estuaries have low %SOU compared to shelves (Fig.5), where the reactivity of organic matter (k_w) is likely similar if not higher because the organic matter is more planktonic-origin. Likewise, low spatial reactivity in shelf waters may be due to higher particle settling velocity for reasons we can only speculate: salinity increase (Abolfazli & Strom, 2023) and blooming of typically dominated elongated or chains-forming phytoplankton species can promote flocculation and fast settling (Arguedas-Leiva et al., 2022). Sediment recycling efficiency regulates %SOC but to a lesser extent (the shaded area in Fig.5), and the effect becomes smaller when the organic matter spatial reactivity is high (see variability in the width of the shaded area in Fig. 5), as there is less organic matter reaching the sediment for recycling efficiency to make a difference.

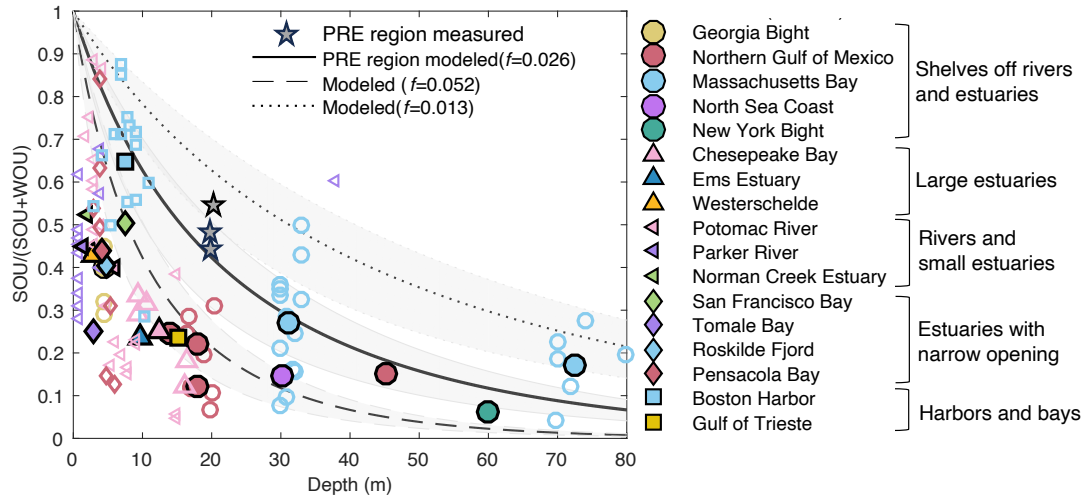


Figure 5 The contributions of SOU to total oxygen uptake (SOU +WOU) as a function of water layer thickness (for the entire water column or BBL). The solid line represents model results for the PRE region ($\varepsilon = 0.5$; $f = 0.026$). Results obtained using different spatial reactivities (f) are shown in dotted and dashed lines; shaded areas show variability with different sediment recycling efficiency ($\varepsilon = 0.3-0.75$). Literature data are from (Boynton et al., 2018; Chi et al., 2021; Kemp et al., 1992; Murrell & Lehrter, 2011). Solid markers indicate average values and the open markers of the corresponding colors are individual measurements.

6. Summary, caveats, and outlooks

Using data in the PRE region, we construct and test a simple mass balance model to understand bottom hypoxia in coastal waters. The model describes oxygen utilization in the BBL due to the sediment and water oxygen uptake (Eqs.8 and 9). While SOU needs to be measured, we show that bottom oxygen levels and water-column stratification can be used to parameterize SOU and increase the data size (Fig.3). WOU is linked to SOU in the model via their intrinsic mass-balance connection (Eqs.7.2 and 10). Therefore, by having data on the water column physical and oxygen conditions across a region (easily obtained from CTD) and some estimates of SOU, the model can estimate other determinants of oxygen loss and quantitatively describe the conditions and time scales for hypoxia formation, as well as the contributions from water column versus sediments.

There are several caveats to using the model. It should be used within areas with similar particle settling velocity and organic matter reactivity. These parameters can vary significantly across regions. Ideally, parameters can be chosen or fitted more locally (e.g., differentiating the inner estuary and the offshore shelf). Moreover, SOU is assumed to be stable during the development of the observed oxygen level, but SOU can decrease when oxygen becomes very limited ($<1.5-3.2 \text{ mg L}^{-1}$) (Chi et al., 2021; Murrell

& Lehrter, 2011; Rowe et al., 2002). Therefore, for the model to work, SOU might need to be measured under sufficient overlying-water oxygen (reaction not limited by oxygen) at sites with very low oxygen levels. Other complications of the physical conditions need to be considered, for example, when stratification is disrupted by a typhoon. Despite these limitations, the model provides intuitive and quantitative estimates of hypoxia and can be used to understand the variability among systems and changes under future scenarios (e.g., longer stratification under climate change). The model can also help estimate key constraints, such as the reactivity and settling velocity of particles, to support other mechanistic models.

Acknowledgment

The work is substantially supported by grants from the Research Grant Council (RGC) of the Hong Kong Special Administrative Region, China (Project Reference Numbers: 16303022 and 26305621 to JL, a grant from the National Natural Science Foundation of China to LY (Project No. 42206160), and grants from RGC to JG and LY (Project Reference Numbers: 16307423 and AoE/P-601/23-N to JG). The work is also funded by the Center for Ocean Research in Hong Kong and Macau (CORE). CORE is a joint research center for ocean research between Laoshan Laboratory and Technology and the Hong Kong University of Science and Technology (HKUST). We thank the captain and crew of the R/V *Haike 68* and Chief Scientists Zhongming Lu and Isaac Cheung for assisting in sample collection. Lei Zhou is acknowledged for assistance in sediment sample collection and field analysis. We thank Mark J. McCarthy (Estonian University of Life Sciences) and Wayne S. Gardner (University of Texas at Austin) for providing CTD data from the Northern Gulf of Mexico. These data were collected aboard the R/V Pelican (Louisiana University Marine Consortium) supported by the NOAA/CSCOR Grant (#NA07NOS4780225) to Wayne Gardner. We thank Weicong Cheng and Dou Li for their technical and intellectual support.

Author contributions:

Conceptualization: Jiying Li

Investigation: Jing Sun, Jiying Li, and Xingyu Yang

Data curation: Jing Sun, Jiying Li, and Xingyu Yang

Formal analysis: Jing Sun, Jiying Li, and Liuqian Yu

Methodology: Jing Sun, Jiying Li, Liuqian Yu, and Jianping Gan

Funding acquisition: Jiying Li, Liuqian Yu, and Jianping Gan

Project administration: Jiying Li and Jianping Gan

Supervision: Jiying Li

Writing—original draft: Jing Sun and Jiying Li

Writing—review& editing: Jiying Li, Liuqian Yu, Hongbin Yin, and Jianping Gan

Data Availability Statement

Data presented in the paper are available at DataSpace@HKUST via doi:10.14711/dataset/KCLUQW (Li et al. 2024). Data from the Changjiang estuary region are from Zhang et al. 2017 and Chi et al. 2021; Data from the Northern Gulf of Mexico are from McCarthy et al. 2013; Data from the Chesapeake Bay are from Boynton et al. (2018), Boynton et al. (2022), Boynton and Ceballos (2019), Kemp et al. (1992), and the Chesapeake Bay Program Datahub (<https://datahub.chesapeakebay.net>). Additional data from other systems are also compiled in Fennel and Testa (2019), Boynton et al. (2018), Boynton et al. (2022), and Kemp et al. (1992).

References

- Abolfazli, E., & Strom, K. (2023). Salinity Impacts on Floc Size and Growth Rate With and Without Natural Organic Matter. *Journal of Geophysical Research: Oceans*, 128(7). <https://doi.org/10.1029/2022jc019255>
- Archer, D., & Devol, A. (1992). Benthic oxygen fluxes on the Washington shelf and slope: A comparison of in situ microelectrode and chamber flux measurements. *Limnology and Oceanography*, 37(3), 614–629. <https://doi.org/10.4319/lo.1992.37.3.0614>
- Arguedas-Leiva, J.-A., Słomka, J., Lalescu, C. C., Stocker, R., & Wilczek, M. (2022). Elongation enhances encounter rates between phytoplankton in turbulence. *Proceedings of the National Academy of Sciences*, 119(32), e2203191119. <https://doi.org/10.1073/pnas.2203191119>
- Bianchi, T. S., DiMarco, S. F., Cowan, J. H., Hetland, R. D., Chapman, P., Day, J. W., & Allison, M. A. (2010). The science of hypoxia in the Northern Gulf of Mexico: A review. *Science of The Total Environment*, 408(7), 1471–1484. <https://doi.org/10.1016/j.scitotenv.2009.11.047>
- Bowman, G. T., & Delfino, J. J. (1980). Sediment oxygen demand techniques: A review and comparison of laboratory and in situ systems. *Water Research*, 14(5), 491–499. [https://doi.org/10.1016/0043-1354\(80\)90215-8](https://doi.org/10.1016/0043-1354(80)90215-8)
- Boynton, Fraser, W. C., & Alejandra, M. (2019). Chesapeake Bay and Maryland Coastal Bays Sediment-Water Oxygen and Nutrient Flux Data Set. *Mendeley Data*, (V1). <https://doi.org/10.17632/jpwvc5jytk.1>
- Boynton, W. R., Ceballos, M. A. C., Bailey, E. M., Hodgkins, C. L. S., Humphrey, J. L., & Testa, J. M. (2018). Oxygen and Nutrient Exchanges at the Sediment-Water Interface: a Global Synthesis and Critique of Estuarine and Coastal Data. *Estuaries and Coasts*, 41(2), 301–333. <https://doi.org/10.1007/s12237-017-0275-5>
- Boynton, W. R., Ceballos, M. A. C., Hodgkins, C. L. S., Liang, D., & Testa, J. M. (2022). Large-Scale Spatial and Temporal Patterns and Importance of Sediment–Water Oxygen and Nutrient Fluxes in the Chesapeake Bay Region. *Estuaries and Coasts*, 1–20. <https://doi.org/10.1007/s12237-022-01127-0>
- Cai, W.-J., Dai, M., Wang, Y., Zhai, W., Huang, T., Chen, S., et al. (2004). The biogeochemistry of inorganic carbon and nutrients in the Pearl River estuary and the adjacent Northern South China Sea. *Continental Shelf Research*, 24(12), 1301–1319. <https://doi.org/10.1016/j.csr.2004.04.005>

- Chi, L., Song, X., Ding, Y., Yuan, Y., Wang, W., Cao, X., et al. (2021). Heterogeneity of the sediment oxygen demand and its contribution to the hypoxia off the Changjiang estuary and its adjacent waters. *Marine Pollution Bulletin*, 172, 112920. <https://doi.org/10.1016/j.marpolbul.2021.112920>
- Diaz, R. J., & Rosenberg, R. (2008). Spreading Dead Zones and Consequences for Marine Ecosystems. *Science*, 321(5891), 926–929. <https://doi.org/10.1126/science.1156401>
- Dzwonkowski, B., Fournier, S., Reager, J. T., Milroy, S., Park, K., Shiller, A. M., et al. (2018). Tracking sea surface salinity and dissolved oxygen on a river-influenced, seasonally stratified shelf, Mississippi Bight, northern Gulf of Mexico. *Continental Shelf Research*, 169, 25–33. <https://doi.org/10.1016/j.csr.2018.09.009>
- Fennel, K., & Testa, J. M. (2019). Biogeochemical Controls on Coastal Hypoxia. *Annual Review of Marine Science*, 11(1), 105–130. <https://doi.org/10.1146/annurev-marine-010318-095138>
- Fennel, K., Laurent, A., Hetland, R., Justić, D., Ko, D. S., Lehrter, J., et al. (2016). Effects of model physics on hypoxia simulations for the northern Gulf of Mexico: A model intercomparison. *Journal of Geophysical Research: Oceans*, 121(8), 5731–5750. <https://doi.org/10.1002/2015jc011577>
- Glud, R. N. (2008). Oxygen dynamics of marine sediments. *Marine Biology Research*, 4(4), 243–289. <https://doi.org/10.1080/17451000801888726>
- Grégoire, M., Garçon, V., Garcia, H., Breitburg, D., Isensee, K., Oschlies, A., et al. (2021). A Global Ocean Oxygen Database and Atlas for Assessing and Predicting Deoxygenation and Ocean Health in the Open and Coastal Ocean. *Frontiers in Marine Science*, 8, 724913. <https://doi.org/10.3389/fmars.2021.724913>
- Katsev, S., Chaillou, G., Sundby, B., & Mucci, A. (2007). Effects of progressive oxygen depletion on sediment diagenesis and fluxes: A model for the lower St. Lawrence River Estuary. *Limnology and Oceanography*, 52(6), 2555–2568. <https://doi.org/10.4319/lo.2007.52.6.2555>
- Kemp, W., Sampou, P., Garber, J., Turtle, J., & Boynton, W. (1992). Seasonal depletion of oxygen from bottom waters of Chesapeake Bay: roles of benthic and planktonic respiration and physical exchange processes. *Marine Ecology Progress Series*, 85, 137–152. <https://doi.org/10.3354/meps085137>
- Laurent, A., & Fennel, K. (2017). Modeling Coastal Hypoxia, Numerical Simulations of Patterns, Controls and Effects of Dissolved Oxygen Dynamics, 149–171. https://doi.org/10.1007/978-3-319-54571-4_7
- Levin, L. A., Ekau, W., Gooday, A. J., Jorissen, F., Middelburg, J. J., Naqvi, S. W. A., et al. (2009). Effects of natural and human-induced hypoxia on coastal benthos. *Biogeosciences*, 6(10), 2063–2098. <https://doi.org/10.5194/bg-6-2063-2009>
- Li, D., Gan, J., Hui, R., Liu, Z., Yu, L., Lu, Z., & Dai, M. (2020). Vortex and Biogeochemical Dynamics for the Hypoxia Formation Within the Coastal Transition Zone off the Pearl River Estuary. *Journal of Geophysical Research: Oceans*, 125(8). <https://doi.org/10.1029/2020jc016178>

- Li, J., & Katsev, S. (2014). Nitrogen cycling in deeply oxygenated sediments: Results in Lake Superior and implications for marine sediments. *Limnology and Oceanography*, 59(2), 465–481. <https://doi.org/10.4319/lo.2014.59.2.0465>
- Li, J., Crowe, S. A., Miklesh, D., Kistner, M., Canfield, D. E., & Katsev, S. (2012). Carbon mineralization and oxygen dynamics in sediments with deep oxygen penetration, Lake Superior. *Limnology and Oceanography*, 57(6), 1634–1650. <https://doi.org/10.4319/lo.2012.57.6.1634>
- Li, J., Zhang, Y., & Katsev, S. (2018a). Phosphorus recycling in deeply oxygenated sediments in Lake Superior controlled by organic matter mineralization. *Limnology and Oceanography*, 63(3), 1372–1385. <https://doi.org/10.1002/lno.10778>
- Li, J., Brown, E. T., Crowe, S. A., & Katsev, S. (2018b). Sediment geochemistry and contributions to carbon and nutrient cycling in a deep meromictic tropical lake: Lake Malawi (East Africa). *Journal of Great Lakes Research*, 44(6), 1221–1234. <https://doi.org/10.1016/j.jglr.2017.12.001>
- Li, J., Sun, J., Yu, L., Gan, J., & Yang, X. (2024). Sediment oxygen uptake and hypoxia in the Pearl River Estuary Region [Dataset]. DataSpace@HKUST, <https://doi:10.14711/dataset/KCLUQW>
- Manning, A. J. (2004). The Observed Effects of Turbulence on Estuarine Flocculation. *Journal of Coastal Research. Sediment Transport in European Estuarine Environments: Proceedings of the STRAEE Workshop (WINTER 2004):*, (Special Issue No. 41), 90–104. Retrieved from <https://www.jstor.org/stable/25736634>
- McCarthy, M. J., Carini, S. A., Liu, Z., Ostrom, N. E., & Gardner, W. S. (2013). Oxygen consumption in the water column and sediments of the northern Gulf of Mexico hypoxic zone. *Estuarine, Coastal and Shelf Science*, 123, 46–53. <https://doi.org/10.1016/j.ecss.2013.02.019>
- Middelburg, J. J., & Levin, L. A. (2009). Coastal hypoxia and sediment biogeochemistry. *Biogeosciences*, 6(7), 1273–1293. <https://doi.org/10.5194/bg-6-1273-2009>
- Moriarty, J. M., Friedrichs, M. A. M., & Harris, C. K. (2021). Seabed Resuspension in the Chesapeake Bay: Implications for Biogeochemical Cycling and Hypoxia. *Estuaries and Coasts*, 44(1), 103–122. <https://doi.org/10.1007/s12237-020-00763-8>
- Murrell, M. C., & Lehrter, J. C. (2011). Sediment and Lower Water Column Oxygen Consumption in the Seasonally Hypoxic Region of the Louisiana Continental Shelf. *Estuaries and Coasts*, 34(5), 912–924. <https://doi.org/10.1007/s12237-010-9351-9>
- Peña, M. A., Katsev, S., Oguz, T., & Gilbert, D. (2010). Modeling dissolved oxygen dynamics and hypoxia. *Biogeosciences*, 7(3), 933–957. <https://doi.org/10.5194/bg-7-933-2010>
- Rowe, G. T., Kaegi, M. E. C., Morse, J. W., Boland, G. S., & Briones, E. G. E. (2002). Sediment community metabolism associated with continental shelf hypoxia, Northern Gulf of Mexico. *Estuaries*, 25(6), 1097–1106. <https://doi.org/10.1007/bf02692207>
- Ruiz, J., Macías, D., & Peters, F. (2004). Turbulence increases the average settling velocity of phytoplankton cells. *Proceedings of the National Academy of Sciences*, 101(51), 17720–17724. <https://doi.org/10.1073/pnas.0401539101>

- Testa, J. M., Li, Y., Lee, Y. J., Li, M., Brady, D. C., Toro, D. M. D., & Kemp, W. M. (2017). Modeling Coastal Hypoxia, Numerical Simulations of Patterns, Controls and Effects of Dissolved Oxygen Dynamics, 95–118. https://doi.org/10.1007/978-3-319-54571-4_5
- Trowbridge, J. H., & Lentz, S. J. (2018). The Bottom Boundary Layer. *Annual Review of Marine Science*, 10(1), 397–420. <https://doi.org/10.1146/annurev-marine-121916-063351>
- Wulff, F., & Stigebrandt, A. (1989). A time-dependent budget model for nutrients in the Baltic Sea. *Glob. Biogeochem. Cycles* 3: 63–78. <https://doi.org/10.1029/gb003i001p00063>
- Yu, L., Fennel, K., Laurent, A., Murrell, M. C., & Lehrter, J. C. (2015a). Numerical analysis of the primary processes controlling oxygen dynamics on the Louisiana shelf. *Biogeosciences*, 12(7), 2063–2076. <https://doi.org/10.5194/bg-12-2063-2015>
- Yu, Liuqian, Fennel, K., & Laurent, A. (2015b). A modeling study of physical controls on hypoxia generation in the northern Gulf of Mexico: Physical Controls on Hypoxia. *Journal of Geophysical Research: Oceans*, 120(7), 5019–5039. <https://doi.org/10.1002/2014jc010634>
- Yu, Liuqian, Gan, J., Dai, M., Hui, C. R., Lu, Z., & Li, D. (2021). Modeling the role of riverine organic matter in hypoxia formation within the coastal transition zone off the Pearl River Estuary. *Limnology and Oceanography*, 66(2), 452–468. <https://doi.org/10.1002/lno.11616>
- Zhang, Haiyan, Zhao, L., Sun, Y., Wang, J., & Wei, H. (2017). Contribution of sediment oxygen demand to hypoxia development off the Changjiang Estuary. *Estuarine, Coastal and Shelf Science*, 192, 149–157. <https://doi.org/10.1016/j.ecss.2017.05.006>
- Zhang, Heng, & Li, S. (2010). Effects of physical and biochemical processes on the dissolved oxygen budget for the Pearl River Estuary during summer. *Journal of Marine Systems*, 79(1–2), 65–88. <https://doi.org/10.1016/j.jmarsys.2009.07.002>
- Zhao, Y., Uthairan, K., Lu, Z., Li, Y., Liu, J., Liu, H., et al. (2021). Destruction and reinstatement of coastal hypoxia in the South China Sea off the Pearl River estuary. *Biogeosciences*, 18(8), 2755–2775. <https://doi.org/10.5194/bg-18-2755-2021>
- Zhou, L. (2022). Sediment oxygen uptake and carbon mineralization in the Pearl River Estuary and adjacent coastal waters. <https://doi.org/10.14711/thesis-991013088359403412>
- Zhu, J., Zhu, Z., Lin, J., Wu, H., & Zhang, J. (2016). Distribution of hypoxia and pycnocline off the Changjiang Estuary, China. *Journal of Marine Systems*, 154, 28–40. <https://doi.org/10.1016/j.jmarsys.2015.05.002>

References From the Supporting Information

- APHA (1998) Standard Methods for the Examination of Water and Wastewater. 20th Edition, American Public Health Association, American Water Works Association and Water Environmental Federation, Washington DC.
- Boynton, W. R., Ceballos, M. A. C., Hodgkins, C. L. S., Liang, D., & Testa, J. M. (2022). Large-Scale Spatial and Temporal Patterns and Importance of Sediment–Water Oxygen and Nutrient Fluxes in the Chesapeake Bay Region. *Estuaries and Coasts*, 1–20. <https://doi.org/10.1007/s12237-022-01127-0>
- Cai, W. J., Dai, M., Wang, Y., Zhai, W., Huang, T., Chen, S., et al. (2004). The biogeochemistry of inorganic carbon and nutrients in the Pearl River estuary and the adjacent Northern South

552 China Sea. *Continental Shelf Research*, 24(12), 1301–1319.
 553 <https://doi.org/10.1016/j.csr.2004.04.005>

554 Cannaby, H., Fach, B. A., Arkin, S. S., & Salihoglu, B. (2015). Climatic controls on biophysical
 555 interactions in the Black Sea under present day conditions and a potential future (A1B) climate
 556 scenario. *Journal of Marine Systems*, 141, 149–166.
 557 <https://doi.org/10.1016/j.jmarsys.2014.08.005>

558 Capet, A., Beckers, J.-M., & Grégoire, M. (2013). Drivers, mechanisms and long-term variability
 559 of seasonal hypoxia on the Black Sea northwestern shelf – is there any recovery after
 560 eutrophication? *Biogeosciences*, 10(6), 3943–3962. <https://doi.org/10.5194/bg-10-3943-2013>

561 Capet, Arthur, Meysman, F. J. R., Akoumianaki, I., Soetaert, K., & Grégoire, M. (2016).
 562 Integrating sediment biogeochemistry into 3D oceanic models: A study of benthic-pelagic
 563 coupling in the Black Sea. *Ocean Modelling*, 101, 83–100.
 564 <https://doi.org/10.1016/j.ocemod.2016.03.006>

565 Fennel, K., & Testa, J. M. (2019). Biogeochemical Controls on Coastal Hypoxia. *Annual Review*
 566 *of Marine Science*. Retrieved from <https://doi.org/10.1146/annurev-marine-010318-095138>

567 Glud, R. N. (2008). Oxygen dynamics of marine sediments. *Marine Biology Research*, 4(4), 243–
 568 289. <https://doi.org/10.1080/17451000801888726>

569 Jørgensen, B. B., & Revsbech, N. P. (1985). Diffusive boundary layers and the oxygen uptake of
 570 sediments and detritus1. *Limnology and Oceanography*, 30(1), 111–122.
 571 <https://doi.org/10.4319/lo.1985.30.1.0111>

572 Lehmann, M. F., Barnett, B., Gélina, Y., Gilbert, D., Maranger, R. J., Mucci, A., et al. (2009).
 573 Aerobic respiration and hypoxia in the Lower St. Lawrence Estuary: Stable isotope ratios of
 574 dissolved oxygen constrain oxygen sink partitioning. *Limnology and Oceanography*, 54(6),
 575 2157–2169. <https://doi.org/10.4319/lo.2009.54.6.2157>

576 Li, J., Crowe, S. A., Miklesh, D., Kistner, M., Canfield, D. E., & Katsev, S. (2012). Carbon
 577 mineralization and oxygen dynamics in sediments with deep oxygen penetration, Lake
 578 Superior. *Limnology and Oceanography*, 57(6), 1634–1650.
 579 <https://doi.org/10.4319/lo.2012.57.6.1634>

580 McCarthy, M. J., Carini, S. A., Liu, Z., Ostrom, N. E., & Gardner, W. S. (2013a). Oxygen
 581 consumption in the water column and sediments of the northern Gulf of Mexico hypoxic zone.
 582 *Estuarine, Coastal and Shelf Science*, 123, 46–53. <https://doi.org/10.1016/j.ecss.2013.02.019>

583 McCarthy, M. J., Carini, S. A., Liu, Z., Ostrom, N. E., & Gardner, W. S. (2013b). Oxygen
 584 consumption in the water column and sediments of the northern Gulf of Mexico hypoxic zone.
 585 *Estuarine, Coastal and Shelf Science*, 123, 46–53. <https://doi.org/10.1016/j.ecss.2013.02.019>

586 McDougall, T. J., & Barker, P. M. (2011). *Getting started with TEOS-10 and the Gibbs Seawater*
 587 *(GSW) Oceanographic Toolbox* (Vol. SCOR/IAPSO WG127). SCOR/IAPSO WG127.

588 Middelburg, J. J. (1989). A simple rate model for organic matter decomposition in marine
 589 sediments. *Geochimica et Cosmochimica Acta*, 53(7), 1577–1581.
 590 [https://doi.org/10.1016/0016-7037\(89\)90239-1](https://doi.org/10.1016/0016-7037(89)90239-1)

591 Noffke, A., Sommer, S., Dale, A. W., Hall, P. O. J., & Pfannkuche, O. (2016). Benthic nutrient
 592 fluxes in the Eastern Gotland Basin (Baltic Sea) with particular focus on microbial mat

ecosystems. *Journal of Marine Systems*, 158, 1–12.
<https://doi.org/10.1016/j.jmarsys.2016.01.007>
 Welsh, B. L., & Eller, F. C. (1991). Mechanisms controlling summertime oxygen depletion in western Long Island Sound. *Estuaries*, 14(3), 265–278. <https://doi.org/10.2307/1351661>
 Wulff, F., & Stigebrandt, A. (1989). A time-dependent budget model for nutrients in the Baltic Sea. *Global Biogeochemical Cycles*, 3(1), 63–78. <https://doi.org/10.1029/gb003i001p00063>
 Zhang, H., Zhao, L., Sun, Y., Wang, J., & Wei, H. (2017). Contribution of sediment oxygen demand to hypoxia development off the Changjiang Estuary. *Estuarine, Coastal and Shelf Science*, 192, 149–157. <https://doi.org/10.1016/j.ecss.2017.05.006>
 Zhou, L. (2022). *Sediment oxygen uptake and carbon mineralization in the Pearl River Estuary and adjacent coastal waters*. Master's Thesis, The Hong Kong University of Science and Technology. doi: 10.14711/thesis-991013088359403412.